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Original Article



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Outdoor $PM_{2.5}$ and their water-soluble ions in the Northern part of the Persian Gulf

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Abstract

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Gulf to the occurrence of dust storms. Outdoor PM_{2.5} and their water-soluble ions in Bushehr port were studied from December 2016 to September 2017. **Methods:** A total of 46 outdoor PM_{2.5} samples were collected by high-volume air sampler and eight water-soluble ions, including Ca²⁺, Mg²⁺, Na⁺, K⁺, F⁻, Cl⁻, NO₃⁻, and SO₄⁻²⁻ in PM_{2.5} were also measured

Background: The environmental conditions potentially predispose the northern part of the Persian

by ion chromatography (IC). **Results:** The 24-hour average concentration of $PM_{2.5}$ was in the range of 22.09 to 292.45 µg/m³. The mean concentration of water-soluble ions in $PM_{2.5}$ was in the range of 0.10 ± 0.14 to 6.76 ± 4.63 µg/m³. The major water-soluble ions were the secondary inorganic aerosols (SO_4^{-2} and NO_3^{-1}), which accounted for nearly 41% of total water-soluble ions in $PM_{2.5}$. The total water-soluble ions level of $PM_{2.5}$ in winter was higher than that in spring and summer. The positive matrix factorization (PMF) model showed that the source contributions of $PM_{2.5}$ were in the order of dust (55.8%), sea salt (17.1%), secondary sulfate (11.8%), industries (7%), vehicular emission (4.7%), and secondary nitrate (3.7%).

Conclusion: According to the results, dust and sea salt are the main sources of water-soluble ions in $PM_{2,5}$ in Bushehr port, which should attract much attention.

Keywords: Aerosols, Bushehr, PM_{2.5}, Seasonal variation, Water-soluble ions

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Introduction

Particulate matter (PM) is one of the air pollutants that is emitted into the atmosphere from different primary or secondary, natural or anthropogenic sources (1,2). Lately, the PM in outdoor air has been categorized as carcinogenic to humans (IARC Group 1) by the International Agency for Research on Cancer (IARC) (3). Outdoor PM can enter the human body via inhalation and ingestion and cause adverse health effects such as lung cancer, atherosclerosis, and respiratory and cardiovascular diseases (4-8). Particle size, surface area, and chemical composition are among the properties of PM that can influence the health effects (9,10).

PM with an aerodynamic diameter smaller than 2.5 μ m (PM_{2.5}) can penetrate deeply into the lung area and deposit on the respiratory tract because of their diameter size (11-13). PM_{2.5} may be generated and emitted into outdoor air from different sources such as power plants,

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waste incineration, crustal and sea salt, vehicle emission construction, agricultural dust, and industrial processes (14,15). $PM_{2.5}$ emitted directly from a source into the atmosphere are called primary $PM_{2.5}$ and some process such as the oxidation of primary gases including, nitrogen oxides, ammonia, volatile organic compounds (VOCs) and sulfur dioxide can produce secondary particles (14,16). Outdoor $PM_{2.5}$ consists of various chemical compounds such as metals, organic carbons, polycyclic aromatic hydrocarbons, and organic and inorganic ions (17-20). It was reported that nitrate and sulfate may be carried by $PM_{2.5}$ and cause various effects including the formation of reactive oxygen species (ROS), cardiovascular disease as well as oxidative stress (21-24).

Bushehr port with its hot and humid weather is located in southwestern Iran along the northern side of the Persian Gulf. Because of the existence of oil and gas resources as well as related petrochemical industries in Bushehr province, this area is considered a sensitive area to air pollution (25,26). Dust storms may also transfer PM and other pollutants from the Arabian Desert located in the southern part of the Persian Gulf. But to date, there is no information on the water-soluble ion content of PM_{2.5} in this part of the Persian Gulf. Therefore, the aims of this study were: (1) to determine water-soluble ions contents of PM_{2.5} including Ca²⁺, Mg²⁺, Na⁺, K⁺, F⁻, Cl⁻, NO₃⁻, and SO₄²⁻, (2) to discuss seasonal variations of water-soluble ions contents in PM_{2.5}, and (3) to identify the source of PM_{2.5} using positive matrix factorization (PMF) model.

Materials and Methods

Study area and PM₂₅ sampling

This study was performed to investigate the water-soluble ion characteristics of $PM_{2.5}$ in Bushehr port. The sampling station was located in Bushehr University of Medical

Sciences building. Figure 1 shows the location of the sampling station in Bushehr port. The outdoor 24-hour samples were collected every six days from December 2016 to September 2017 for 9 months. During the sampling period, 46 samples were collected for investigating water-soluble ions of PM_{2.5}. Samples were collected on quartz fiber filters (8 in × 10 in, USA, pore size: 2.5 μ m) using a PM_{2.5} Tisch high-volume sampler (Model TE-6070D, USA) at a flow rate of 1.42 to 1.58 m³/min. The meteorological data including temperature, relative humidity, and wind speed were collected from meteorological stations near the sampling station in Bushehr port during the sampling period. The range of temperature, relative humidity, and wind speed were 8.12–42.2° C, 31.93–78.12%, and 1.62–9.25 m/s, respectively.

Chemical analysis

Before and after sampling, each filter was kept in a dehydrated desiccator for 48 h under a relative humidity of 25-30% and temperature of 20-25 °C (27,28). To specify $PM_{2.5}$ mass concentration, the filters were weighted three times before and after the sampling using an electronic microbalance with a sensitivity of 0.1 mg. The $PM_{2.5}$ mass concentration was estimated using Eq. (1):

$$PM_{2.5}(\mu g/m^3) = \frac{W_f - W_i \times 10^6}{V}$$
(1)

Where W_f and W_i are the weights of the filter before and after the sampling (g), V represents the volume of the air sample (m³), and 10⁶ is a factor for converting grams to micrograms.

After weighing, the filters containing $PM_{2.5}$ particles were packed and reserved at $-18^{\circ}C$ until the extraction and chemical analysis. Then, the filters were sliced into four equal fractions and one-fourth of the filter was used

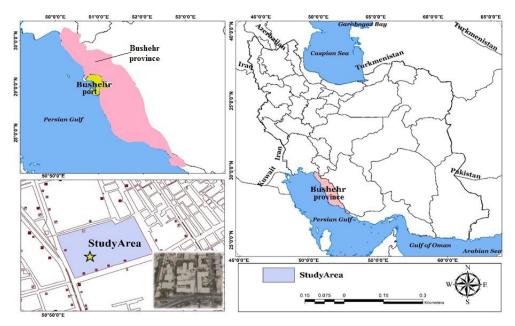


Figure 1. The map of the study area represents the sampling station (17)

to determine water-soluble ions (Ca2+, Mg2+, Na+, K+, F-, Cl^{-} , NO_{3}^{-} , SO_{4}^{-}) contents of PM_{25} particles. One-fourth of each sample filter was sliced and all fragments were added to a glass vial and 50 ml of ultrapure water was added. Then, the vial was shaken for 60 minutes and after that was ultrasonicated for 30 min to complete extraction. The extracted solution was filtered using a microporous membrane (pore size: 0.45 µm and diameter: 25 mm), and the solution was reserved in plastic vials at -4 °C until chemical analysis (29). The water-soluble ions including Ca²⁺, Mg²⁺, Na⁺, K⁺, F⁻, Cl⁻, NO₃⁻, and SO₄⁻² were analyzed using an ion chromatograph (Metrohm 850 Professional IC, Switzerland). The data quantification process was performed using an external standard method. The standard curves were graphed according to the fiveconcentration gradient of the standard. The correlation coefficient (r²) of calibration curves for each ion species was higher than 0.99.

*PM*_{2,5} source apportionment

The PMF is a receptor model developed by the Environmental Protection Agency (EPA), which is used to reconstruct the contribution of emissions from different sources of atmospheric pollutants (e.g., PM_{2,5}) based on ambient measurement data (i.e., PM₂₅ chemical composition). The latest version of PMF 5.0 decomposes a matrix of spectate sample data into factor contributions and factor profiles (30,31). In this study, PM₂₅ sources during the sample collection period were quantified by the model. In total, eight ion species (Ca²⁺, Mg²⁺, Na⁺, K⁺, F, Cl, NO₃, and SO₄²⁻) were applied in the PMF model. It should be noted that PMF was created to run in robust mode and eight-factor solution was achieved from the base run. Both Q robust and Q true values were found to be in a fair agreement that this implicates the stability of the model and proficiency to reasonably fit all the variables.

Statistical analysis

The SPSS 19 software and Microsoft Excel 2016 were used for statistical analysis. The descriptive statistics and the spatial-temporal distribution pattern were applied to analyze $PM_{2.5}$ and its water-soluble ions contents. Furthermore, Pearson's correlation analysis was employed to identify the relationships between $PM_{2.5}$ and ionic species. Statistical significance was accepted at P < 0.05 and P < 0.01.

Results

PM_{25} mass concentration

The PM_{2.5} concentration ranged from 22.09 to 292.45 μ g/m³ with a mean concentration of 65.77 μ g/m³. During the sampling period, just 2 samples were lower than the WHO guideline (32) and 8 samples were lower than the EPA guideline (33). The highest concentration of PM_{2.5}

 $(292.44 \,\mu\text{g/m}^3)$ was reported in April (19,34).

Mass concentration of water-soluble ions

The descriptive statistics for $PM_{2.5}$ water-soluble ions $(Ca^{2+}, Mg^{2+}, Na^+, K^+, F^-, Cl^-, NO_3^-, and SO_4^{2-})$ during the sampling period are presented in Table 1. The concentrations of water-soluble ions including Ca^{2+} , Mg^{2+} , Na^+ , K^+ , F^- , Cl^- , NO_3^- , and SO_4^{-2-} were in the range of 0.97–8.43, 0.07–0.64, 4.29–8.12, 2.63–4.3, 0.02–0.84, 0.004–4.58, 0.004–22.77, and 0.02–23.55 µg/m³, respectively.

Figure 2 shows the contribution of each watersoluble ions to the total mass of water-soluble ions and $PM_{2.5}$, respectively. As shown in this figure, the order of contribution of water-soluble ions to the total mass of ions as well as $PM_{2.5}$ was as: $SO_4^{2-} > Na^+ > NO_3^{-} > Ca^{2+} > K^+ > Cl^ > Mg^{2+} > F^-$. The ions SO_4^{2-} , Na^+ , NO_3^{-} , Ca^{2+} , and K^+ were the most abundant water-soluble ions in the study area.

Temporal and seasonal trends of water-soluble ions in $PM_{2.5}$

Temporal and seasonal variations of the water-soluble ions (Ca²⁺, Mg²⁺, Na⁺, K⁺, F⁻, Cl⁻, NO₃⁻, and SO₄²⁻) in PM_{2.5} are shown in Figures 3 and 4. As can be seen in Figure 3, no specific temporal trends were observed for any of the water-soluble ions. Almost the concentrations of all water-soluble ions except Na⁺and K⁺had a sharp peak during the sampling period. Six components of water-soluble ions including Ca²⁺, Na⁺, K⁺, Cl⁻, NO₃⁻, and SO₄²⁻ had a concentration peak in dusty days during the sampling period.

Bivariate correlations were performed to assess the relationships between $PM_{2.5}$ and ionic species as shown in Table 2. The correlation analysis indicated that $PM_{2.5}$ level was positively correlated with NO_3^- and Ca^{2+} . Na⁺showed a positive correlation with NO_3^- , Ca^{2+} , and K⁺. Also, a positive correlation was observed between K⁺and Cl⁻. SO₄²⁻ illustrated a negative relationship with F⁻ and a positive correlation with Cl⁻. Moreover, Cl⁻had a positive correlation with NO₃⁻, SO₄²⁻, and K⁺. Ca²⁺was positively related to NO_3^- and Mg^{2+} .

Table 1. Summary statistics of $\mathsf{PM}_{2.5}$ and their water-soluble ions contents $(\mu g/m^3)$

(µg/m)						
Species	Mean	SD	Min	Мах	Median	IQR
PM _{2.5}	65.77	49.84	22.09	292.45	52.8	29.13
Ca ²⁺	3.68	1.33	0.97	8.43	3.49	1.25
Mg ²⁺	0.28	0.14	0.07	0.64	0.27	0.25
Na⁺	5.38	1.01	4.29	8.12	4.82	1.4
K⁺	3.55	0.43	2.63	4.3	3.62	0.55
F	0.10	0.14	0.02	0.84	0.05	0.05
Cl-	2.06	1.21	0.004	4.58	2.33	1.26
NO ₃ ⁻	3.88	3.70	0.004	22.77	3.52	3.05
SO42-	6.76	4.63	0.02	23.55	6.27	4.59

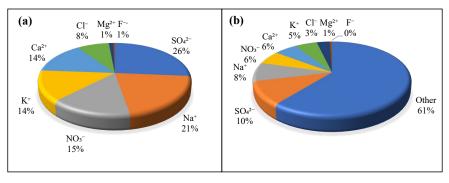
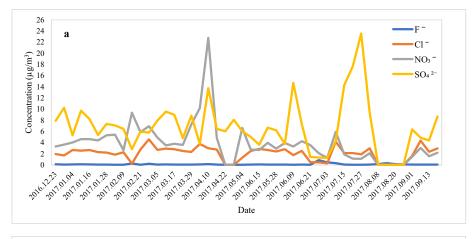


Figure 2. Relative contribution of water-soluble ions to the total mass of ions (a) and PM₂₅(b)



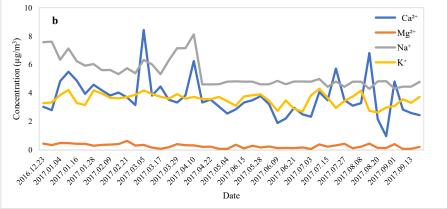
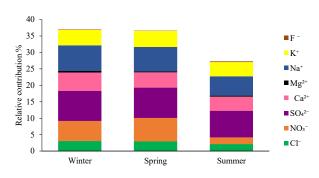


Figure 3. Temporal variations of the water-soluble ions, a: Anions and b: Cations



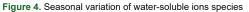


Table 2. Correlation coefficient among PM₂₅ and ionic species

	2.5								
	PM _{2.5}	F	CI.	NO ₃ .	SO42-	Ca ²⁺	Mg ²⁺	Na⁺	K⁺
$PM_{_{2.5}}$	1								
F⁻	-0.002	1							
Cl-	0.14	-0.26	1						
NO ₃ -	0.54**	01	0.41**	1					
SO42-	0.19	-0.37 [*]	0.3*	0.17	1				
Ca ²⁺	0.51 [™]	-0.05	0.18	0.32*	0.21	1			
Mg ²⁺	0.06	-0.04	0.16	0.25	0.18	0.51**	1		
Na⁺	0.25	-0.09	0.27	0.62**	0.17	0.34*	0.46**	1	
K⁺	0.24	-0.19	0.43**	0.28	0.09	0.29	0.13	0.28	1
*Correlation is significant at the 0.01 level.									

*Correlation is significant at the 0.05 level.

Source apportionment by PMF model

The sources of Ca²⁺, Mg²⁺, Na⁺, K⁺, F⁻, Cl⁻, NO₃⁻, and SO₄²⁻ in PM_{2.5} were quantified by PMF 5.0 model, and six factors including secondary sulfate (factor 1), dust (factor 2), industries (factor 3), vehicular emission (factor 4), secondary nitrate (factor 5), and sea salt (factor 6) were identified. The sources of PM_{2.5} for the six factors were 11.8%, 55.8%, 7%, 4.7%, 3.7%, and 17.1%, respectively (Figure 5). The factors profile and the percentage of PM_{2.5} and species apportion to each source are presented in Figures 6 and 7, respectively.

Discussion

Water-soluble ions in PM_{2,5} and temporal trends

According to Figure 2, SO_4^{2-} , Na^+ , NO_3^- , Ca^{2+} , and K⁺accounted for approximately 90% and 35% of the total mass of water-soluble ions and $PM_{2.5}$, respectively. Shahsavani et al (35) reported that ionic components in Ahvaz, Iran, were in order of $SO_4^{-2} > Ca^{2+} > NO_3^{-} > Mg^{2+} > Cl^- > Na^+ \sim NH_4^{+} > K^+ > F^- > NO_2^{-1}$ in TSP and $Ca^{2+} > NO_3^{-} > SO_4^{-2-} > Mg^{2+} > Cl^- > Na^+ \sim NH_4^{+} > K^+ > F^- > NO_2^{-1}$ in PM_{10} . Also, Ca^{2+} , Mg^{2+} , NO_3^{-7} , SO_4^{-2-} , and Cl⁻ were respectively the most abundant ions in TSP and PM_{10} . Naimabadi et al (36) reported that the most frequent water-soluble ions in PM_{10} in Ahvaz, Iran, were SO_4^{-2-} , Cl^- , NO_3^{--} , and Ca^{2+} , and contributions of water-soluble ions were in order of $SO_4^{-2-} > NO_3^{--} > Na^+ > Ca^{2+} > Cl^- > K^+ > Mg^{2+}$. Goudarzi et al (37) reported that water-soluble ions in Ahvaz, Iran, were in order of $SO_4^{-2-} > NO_3^{-2-} > Na^+ > Ca^{2+} > Cl^- > K^+ > Mg^{2+}$.

In Table 3, the concentrations of water-soluble ions in the present study are compared with those in other studies in different countries. According to the results, Deng et al (38), Hien et al (39), Sharma et al (40), Khan et al (41), and Remoundaki et al (42) reported SO_4^{2-} as the most abundant ion in PM between all water-soluble ions. In contrast, Zhao et al (43), Hassanvand et al (29), and Saraga et al (44) reported NO_3^{-} as the most abundant ion in PM.

The sea salt species of water-soluble ions (Na⁺and Cl⁻) accounted for almost 29% of the total mass of

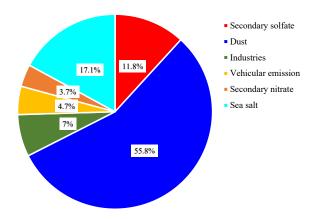


Figure 5. The contribution of each source to the ambient $\mathrm{PM}_{_{2.5}}(\%)$ during the sampling period

water-soluble ions. The high concentrations of these compounds are associated with the proximity of the study area to the Persian Gulf. The contribution of sea salt particles may be a source of high Cl⁻ concentration in many marine environments including the Persian Gulf (46). In agreement with the results of the present study, high concentrations of Na⁺and Cl⁻ were reported by Gholampour et al (47), and Ho et al (48), who reported the high concentrations of these ions can be due to strong and permanent winds that blow on the beach and can create water droplets and aerosols.

Considering Na⁺as a tracer of sea salt, excess sulfate (non-sea-salt sulfates [nss- SO_4^{2-}]), excess potassium (non-sea-salt potassium (nss- K^+)), and excess calcium (non-sea-salt calcium (nss- Ca^{2+})) can be estimated using the following equations (49):

nss - $SO_4^{2-} = [SO_4^{2-}] - [Na^+] \times 0.2516$	(2)
nss - $K^+ = [K^+] - [Na^+] \times 0.037$	(3)
nss - $Ca^{2+} = [Ca^{2+}] - [Na^+] \times 0.0385$	(4)

The non-sea-salt SO_4^{2-} , K⁺, and Ca^{2+} accounted for nearly 80%, 94%, and 92% of the total sulfate, potassium, and calcium of PM_{2.5} that may originate from other sources such as dust, industries, and vehicular emissions.

In agreement with the results of the present study, Yan et al, reported that nss-SO42- in the East China Sea ranged from 1.46-24.9 µg/m3, which accounted for 90.4% of total sulfate (50). In another study in the northern South China Sea, Hsu et al, also reported that nss-SO₄²⁻ in PM₂₅ accounted for 95.9% of total sulfate, indicating anthropogenic sources of sulfate (51). But in a study around Urmia Lake, Iran, Gholampour et al (47) reported that the amounts of excess sulfate and excess potassium were approximately 65%-75% and 60%-75%, respectively, which are lower than the evaluated values in the present study. As Bushehr port is located in a hot and humid area and has numerous industries, higher concentrations of SO_4^{2-} could be because of the increased photochemical oxidation of SO₂ during warm periods, which is consistent with the results of former studies in Thessaloniki and Budapest (52,53). Also, the particulate forms of sulfate and nitrate may be produced due to the rapid oxidation of SO₂ and NOx produced from various sources, especially industries (54,55).

The mass ratio of nitrate/sulfate is a good indicator of the relative importance of stationary and mobile sources of nitrogen as well as sulfur in the atmosphere (56). In this ratio, nitrate and sulfate are used as indicators of mobile and stationary emission, respectively. In the present study, the mass ratio of nitrate to sulfate was 0.57, indicating that stationary emission was the predominant source in the study area. It was expected that stationary emission was the dominant source because there is not much traffic in Bushehr port. Also, the mass ratio of nitrate to sulfate

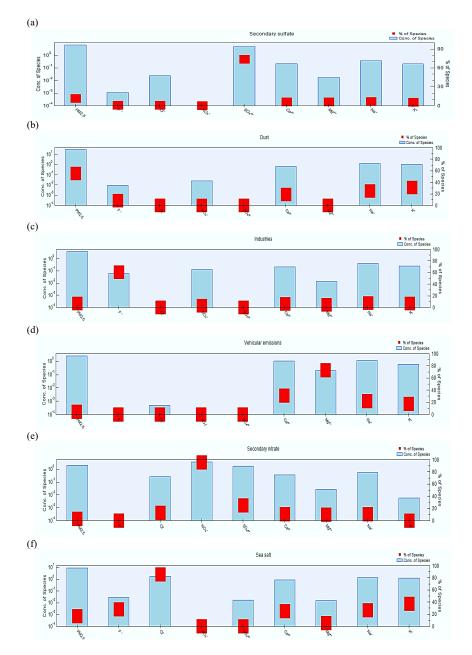


Figure 6. Source profiles of $\mathrm{PM}_{_{25}}$ and other water-soluble ions in the northern part of the Persian Gulf

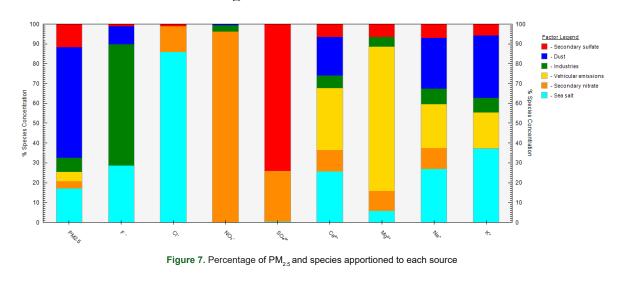


Table 3. Comparative evaluation of water-soluble ions in the present study and former studies (µg/m³)

Country	Site		Year	Ca ²⁺	Mg ²⁺	Na⁺	K⁺	F [.]	Cl.	NO ₃ -	SO42-	Reference
Iran Tehran	T 1	Retirement home	0010 0010	0.97	0.10	0.60	0.06	-	0.47	4.00	3.40	(20)
	School dormitory	2012-2013	1.10	0.50	0.40	0.08	-	0.70	5.70	4.50	(29)	
China	Hefei		2012- 2013	5.24	0.30	0.48	0.96	-	1.21	15.14	15.56	(38)
Vietnam	Hanoi		1999-2001	0.33	0.05	0.14	0.61	-	0.03	0.33	6.47	(39)
India	Delhi		2013-2014	2.83	0.96	5.05	4.10	0.91	7.77	10.0	12.9	(40)
Japan	Yokoham	а	2007-2008	0.20	0.05	0.25	0.13	-	0.21	0.96	3.8	(41)
Greece	Athens		2010-2010	0.34	0.07	0.28	0.10	-	0.23	0.49	3.99	(42)
China	Beijing		2009-2010	1.50	0.20	0.50	1.70	-	2.90	20.50	19.10	(43)
Qatar	Doha		2015	7.42	0.31	1.26	0.23	-	1.4	30.5	15.24	(44)
Iran	Sistan		2014	10.10	0.80	3.14	6.95	-	2.31	1.57	4.06	(45)
Iran	Bushehr		2016-2017	3.68	0.28	5.38	3.55	0.1	2.06	3.88	6.76	This study

in winter, spring, and summer were 0.66, 0.77, and 0.24, respectively. According to the mass ratio of nitrate to sulfate in each season, the stationary emission was the highest in summer followed by winter and spring.

Temporal trends of water-soluble ions are consistent with the reports of previous studies on the ions content of PM₂₅ during dust storms (35,36,46). Also, seasonal variation of total water-soluble ions levels (Figure 4) showed that the highest concentrations of ions were measured in winter (with a mean of $3.64 \,\mu\text{g/m}^3$) followed by spring (with a mean of $3.29 \,\mu\text{g/m}^3$), and summer (with a mean of 2.69 $\mu g/m^3),$ respectively. These results are consistent with the results of previous studies (56-59). According to Figure 4, the concentrations of Ca^{2+} , Mg^{2+} , Na⁺, K⁺, F⁻, and SO^{$^{2}} were slightly different throughout$ </sup> all seasons. Besides the eight species, the concentrations of Cl⁻ and NO₃⁻ were lower than those in summer. The seasonal variation of PM_{2.5} and the consequence of watersoluble ions can be due to atmospheric conditions such as temperature, wind speed, and humidity (58). The concentrations of water-soluble ions are correlated with dust concentration (37,60). Also, the air mass trajectories during the sampling period in Bushehr port can affect the content of PM₂₅. Based on another study in Bushehr port (34) at the same sampling periods and places of the present study, the air mass from the northwest and southeast had the longest trajectory lines, indicating the importance of the west and southwest regions including Iraq, Saudi Arabia, and Khark Island, with high oil/gaslinked activities. Also, the trajectory analysis indicated that water-soluble ions of PM_{2.5} in Bushehr port were originated from distal west regions during the sampling days, especially on dusty days.

The relationships between $PM_{2.5}$ and ionic species are illustrated in Table 2. Usually, an R^2 value of almost 0.5 is considered a good correlation between the cations and anions, showing the probability of the formation of that specific cation and anion (61). The results of a study in Durg City, India, were consistent with the results of this

study and showed a positive correlation between $PM_{2.5}$ and NO_3^- (62). The good correlation between $PM_{2.5}$ and NO_3^- may be due to the occurrence of the component of NaNO₃ in $PM_{2.5}$ (63). In a study by Shahsavani et al, a good correlation was reported between Na⁺and Cl⁻ (0.56), NH₄⁺and SO₄⁻²⁻ (0.49), K⁺and Cl⁻ (0.58), K⁺and SO₄⁻²⁻ (0.56), Ca²⁺and Cl⁻ (0.64), Ca²⁺and NO₃⁻ (0.56), and Ca²⁺with SO₄⁻²⁻ (0.50). They concluded that NaCl, $(NH_4)_2SO_4$, KCl, K₂SO₄, CaCl₂, Ca(NO₃)₂, and CaSO₄ may exist in the total suspended particles (35).

Source apportionment by PMF model

The sources of $PM_{2.5}$ for the six factors including secondary sulfate, dust, industries, vehicular emission, secondary nitrate, and sea salt were identified in Figure 5. Also, Figures 6 and 7 present the factors profile and the percentage of $PM_{2.5}$ and species apportion to each source, respectively.

The first factor, weighted by SO_4^{2-} , is associated with secondary sulfate origin (Figures 6a and 7). Secondary sulfate is typically related to long-range transportation events. Due to the slow oxidation of SO_2 to SO_4^{2-} , this aerosol component is more associated with transportation than local pollution (64), as found in background areas on the Mediterranean coast (65).

The profile of factor 2 is mostly determined by Ca^{2+} , Na^+ , and K^+ (Figures 6b and 7), which are associated with dust storms (66).

Factor 3 was heavily weighted by F⁻ (Figures 6c and 7), which is mostly used as an indicator for industrial sources (67,68).

Factor 4 shows the vehicular emissions source because of the high portions of Ca^{2+} , Mg^{2+} , Na^+ , and K^+ (Figures 6d and 7). This factor relates to vehicular emissions sources due to diesel and gasoline-powered vehicles exhaust emissions (69,70).

Factor 5 indicates high burdens for NO_3^- (Figures 6e and 7), which can be known as a secondary nitrate source. As a result, their gaseous precursors species such as NO_x

emitted from vehicles, fossil, and industry processes can cause the formation of secondary nitrate sources (69).

The sea salt source (factor 6) is heavily weighted by the attendance of Cl⁻ (Figures 6f and 7), indicating the sea spray (71). The presence of sea salt is fully usual in the contaminated coastal and marine atmosphere (72,73), which was also detected in the southeastern part of China in PM_{25} (74).

Conclusion

The present study was performed in the northern part of the Persian Gulf (in Bushehr port) for 9 months from Dec 2016 to Sep 2017 to determine the water-soluble ions contents of PM25 including Ca2+, Mg2+, Na+, K+, F-, Cl-, NO_{3}^{-} , and SO_{4}^{-2} . The mean mass concentration of PM_{25}^{-} during the sampling period was 65.77 µg/m³. Only 4% and 17% of the samples were lower than the daily mean of the WHO and EPA outdoor air guidelines, respectively. Also, the mean concentration of PM_{2.5} on dusty days was nearly 3 times higher than that on other days, indicating that dust storm can be a source of increased PM₂₅ in outdoor air in this area. These water-soluble ions accounted for almost 90% and 35% of the total mass of water-soluble ions and mass of PM2,5, respectively. The water-soluble ions including Ca2+, Na+, K+, Cl-, SO42, and NO_{3}^{-} in PM_{25}^{-} had a concentration peak in dusty days during the sampling period. Also, source apportionment results showed that dust, sea salts, secondary sulfate, industries, vehicular emission, and secondary nitrate were the major contributors to PM_{2.5} mass. Therefore, due to the geographical location of Bushehr province, the occurrence of dust storms and the presence of numerous industries in this area is highly needed to control various sources of PM_{2.5} emissions and choose suitable strategies to reduce the health effects of PM_{25} .

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Authors' contribution

Conceptualization: Sina Dobaradaran. Data curation: Fatemeh Faraji Ghasemi. Formal analysis: Fatemeh Faraji Ghasemi. Funding acquisition: Sina Dobaradaran. Investigation: Sina Dobaradaran. Methodology: Fatemeh Faraji Ghasemi, Sina Dobaradaran, Azam Mohammadi. Project administration: Sina Dobaradaran. Resources: Sina Dobaradaran. Software: Marzieh Mahmoodi. Supervision: Sina Dobaradaran. Validation: Sina Dobaradaran. Visualization: Sina Dobaradaran. Writing-original draft: Fatemeh Faraji Ghasemi. Writing-review & editing: Sina Dobaradaran, Reza Saeedi, Azam Mohammadi, Amir Hossein Darabi, Marzieh Mahmoodi.

Competing interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Ethical issues

The authors certify that all data collected during the study are as stated in the manuscript, and no data from the study has been or will be published separately elsewhere. Ethical approval was obtained from Bushehr University of Medical Sciences (IR.BPUMS.REC.1395.72).

References

- Dumka UC, Tiwari S, Kaskaoutis DG, Hopke PK, Singh J, Srivastava AK, et al. Assessment of PM2.5 chemical compositions in Delhi: primary vs secondary emissions and contribution to light extinction coefficient and visibility degradation. J Atmos Chem. 2017;74(4):423-50. doi: 10.1007/s10874-016-9350-8.
- Natusch DF, Wallace JR, Evans CA Jr. Toxic trace elements: preferential concentration in respirable particles. Science. 1974;183(4121):202-4. doi: 10.1126/science.183.4121.202.
- Loomis D, Grosse Y, Lauby-Secretan B, El Ghissassi F, Bouvard V, Benbrahim-Tallaa L, et al. The carcinogenicity of outdoor air pollution. Lancet Oncol. 2013;14(13):1262-3. doi: 10.1016/s1470-2045(13)70487-x.
- Pope CA 3rd, Dockery DW. Health effects of fine particulate air pollution: lines that connect. J Air Waste Manag Assoc. 2006;56(6):709-42. doi: 10.1080/10473289.2006.10464485.
- Hu X, Zhang Y, Ding Z, Wang T, Lian H, Sun Y, et al. Bioaccessibility and health risk of arsenic and heavy metals (Cd, Co, Cr, Cu, Ni, Pb, Zn and Mn) in TSP and PM2.5 in Nanjing, China. Atmos Environ. 2012;57:146-52. doi: 10.1016/j.atmosenv.2012.04.056.
- Kam W, Cheung K, Daher N, Sioutas C. Particulate matter (PM) concentrations in underground and ground-level rail systems of the Los Angeles Metro. Atmos Environ. 2011;45(8):1506-16. doi: 10.1016/j.atmosenv.2010.12.049.
- Dehghani M, Anushiravani A, Hashemi H, Shamsedini N. Survey on air pollution and cardiopulmonary mortality in shiraz from 2011 to 2012: an analytical-descriptive study. Int J Prev Med. 2014;5(6):734-40.
- Mosallaei S, Hashemi H, Hoseini M, Dehghani M, Naz A. Polycyclic Aromatic Hydrocarbons (PAHs) in household dust: the association between PAHs, cancer risk and sick building syndrome. Build Environ. 2023;229:109966. doi: 10.1016/j.buildenv.2022.109966.
- 9. Kelly FJ, Fussell JC. Size, source and chemical composition as determinants of toxicity attributable to ambient particulate matter. Atmos Environ. 2012;60:504-26. doi:

10.1016/j.atmosenv.2012.06.039.

- Pui DY, Chen SC, Zuo Z. PM2.5 in China: measurements, sources, visibility and health effects, and mitigation. Particuology. 2014;13:1-26. doi: 10.1016/j. partic.2013.11.001.
- Delfino RJ, Sioutas C, Malik S. Potential role of ultrafine particles in associations between airborne particle mass and cardiovascular health. Environ Health Perspect. 2005;113(8):934-46. doi: 10.1289/ehp.7938.
- Putaud JP, Van Dingenen R, Alastuey A, Bauer H, Birmili W, Cyrys J, et al. A European aerosol phenomenology–3: physical and chemical characteristics of particulate matter from 60 rural, urban, and kerbside sites across Europe. Atmos Environ. 2010;44(10):1308-20. doi: 10.1016/j. atmosenv.2009.12.011.
- Utell MJ, Frampton MW. Acute health effects of ambient air pollution: the ultrafine particle hypothesis. J Aerosol Med. 2000;13(4):355-59. doi: 10.1089/jam.2000.13.355.
- Goldstein AH, Koven CD, Heald CL, Fung IY. Biogenic carbon and anthropogenic pollutants combine to form a cooling haze over the southeastern United States. Proc Natl Acad Sci U S A. 2009;106(22):8835-40. doi: 10.1073/ pnas.0904128106.
- Juda-Rezler K, Reizer M, Oudinet JP. Determination and analysis of PM10 source apportionment during episodes of air pollution in Central Eastern European urban areas: the case of wintertime 2006. Atmos Environ. 2011;45(36):6557-66. doi: 10.1016/j.atmosenv.2011.08.020.
- Atkinson RW, Fuller GW, Anderson HR, Harrison RM, Armstrong B. Urban ambient particle metrics and health: a time-series analysis. Epidemiology. 2010;21(4):501-11. doi: 10.1097/EDE.0b013e3181debc88.
- Callén MS, de la Cruz MT, López JM, Navarro MV, Mastral AM. Comparison of receptor models for source apportionment of the PM10 in Zaragoza (Spain). Chemosphere. 2009;76(8):1120-9. doi: 10.1016/j. chemosphere.2009.04.015.
- Kim KH, Jahan SA, Kabir E, Brown RJ. A review of airborne polycyclic aromatic hydrocarbons (PAHs) and their human health effects. Environ Int. 2013;60:71-80. doi: 10.1016/j.envint.2013.07.019.
- Faraji Ghasemi F, Dobaradaran S, Saeedi R, Nabipour I, Nazmara S, Ranjbar Vakil Abadi D, et al. Levels and ecological and health risk assessment of PM2.5-bound heavy metals in the northern part of the Persian Gulf. Environ Sci Pollut Res Int. 2020;27(5):5305-13. doi: 10.1007/s11356-019-07272-7.
- Mirzaei S, Hashemi H, Hoseini M. Concentration and potential source identification of trace elements in wet atmospheric precipitation of Shiraz, Iran. J Environ Health Sci Eng. 2018;16(2):229-37. doi: 10.1007/s40201-018-0310-x.
- Alessandria L, Schilirò T, Degan R, Traversi D, Gilli G. Cytotoxic response in human lung epithelial cells and ion characteristics of urban-air particles from Torino, a northern Italian city. Environ Sci Pollut Res Int. 2014;21(8):5554-64. doi: 10.1007/s11356-013-2468-1.
- 22. Danielsen PH, Loft S, Kocbach A, Schwarze PE, Møller P. Oxidative damage to DNA and repair induced by Norwegian wood smoke particles in human A549 and THP-1 cell lines. Mutat Res. 2009;674(1-2):116-22. doi: 10.1016/j.mrgentox.2008.10.014.
- 23. Ghio AJ, Carraway MS, Madden MC. Composition of air

pollution particles and oxidative stress in cells, tissues, and living systems. J Toxicol Environ Health B Crit Rev. 2012;15(1):1-21. doi: 10.1080/10937404.2012.632359.

- Rezaei M, Salimi A, Taghidust M, Naserzadeh P, Goudarzi G, Seydi E, et al. A comparison of toxicity mechanisms of dust storm particles collected in the southwest of Iran on lung and skin using isolated mitochondria. Toxicol Environ Chem. 2014;96(5):814-30. doi: 10.1080/02772248.2014.959317.
- 25. Gheybi MK, Movahed A, Dehdari R, Amiri S, Khazaei HA, Gooya M, et al. Dusty air pollution is associated with an increased risk of allergic diseases in southwestern part of Iran. Iran J Allergy Asthma Immunol. 2014;13(6):404-11.
- Shabankarehfard E, Ostovar A, Farrokhi S, Naeimi B, Zaeri S, Nazmara S, et al. Air- and dust-borne fungi in indoor and outdoor home of allergic patients in a dust-storm-affected area. Immunol Invest. 2017;46(6):577-89. doi: 10.1080/08820139.2017.1322102.
- Rohra H, Tiwari R, Khare P, Taneja A. Indoor-outdoor association of particulate matter and bounded elemental composition within coarse, quasi-accumulation and quasiultrafine ranges in residential areas of northern India. Sci Total Environ. 2018;631-632:1383-97. doi: 10.1016/j. scitotenv.2018.03.095.
- Taner S, Pekey B, Pekey H. Fine particulate matter in the indoor air of barbeque restaurants: elemental compositions, sources and health risks. Sci Total Environ. 2013;454-455:79-87. doi: 10.1016/j.scitotenv.2013.03.018.
- Hassanvand MS, Naddafi K, Faridi S, Arhami M, Nabizadeh R, Sowlat MH, et al. Indoor/outdoor relationships of PM10, PM2.5, and PM1 mass concentrations and their water-soluble ions in a retirement home and a school dormitory. Atmos Environ. 2014;82:375-82. doi: 10.1016/j. atmosenv.2013.10.048.
- Paatero P. Least squares formulation of robust non-negative factor analysis. Chemometr Intell Lab Syst. 1997;37(1):23-35. doi: 10.1016/s0169-7439(96)00044-5.
- Paatero P, Tapper U. Positive matrix factorization: a nonnegative factor model with optimal utilization of error estimates of data values. Environmetrics. 1994;5(2):111-26. doi: 10.1002/env.3170050203.
- World Health Organization (WHO). Particulate matter, ozone, nitrogen dioxide and sulfur dioxide. In: Air Quality Guidelines: Global Update. WHO; 2005. p. 76-100.
- Esworthy R. 2006 National Ambient Air Quality Standards (NAAQS) for Fine Particulate Matter (PM2.5): Designating Nonattainment Areas. Library of Congress, Congressional Research Service; 2012.
- 34. Akhbarizadeh R, Dobaradaran S, Amouei Torkmahalleh M, Saeedi R, Aibaghi R, Faraji Ghasemi F. Suspended fine particulate matter (PM2.5), microplastics (MPs), and polycyclic aromatic hydrocarbons (PAHs) in air: their possible relationships and health implications. Environ Res. 2021;192:110339. doi: 10.1016/j.envres.2020.110339.
- 35. Shahsavani A, Naddafi K, Jaafarzadeh Haghighifard N, Mesdaghinia A, Yunesian M, Nabizadeh R, et al. Characterization of ionic composition of TSP and PM10 during the Middle Eastern Dust (MED) storms in Ahvaz, Iran. Environ Monit Assess. 2012;184(11):6683-92. doi: 10.1007/s10661-011-2451-6.
- 36. Naimabadi A, Ghadiri A, Idani E, Babaei AA, Alavi N, Shirmardi M, et al. Chemical composition of PM10 and its in vitro toxicological impacts on lung cells during the

Middle Eastern Dust (MED) storms in Ahvaz, Iran. Environ Pollut. 2016;211:316-24. doi: 10.1016/j.envpol.2016.01.006.

- 37. Goudarzi G, Shirmardi M, Naimabadi A, Ghadiri A, Sajedifar J. Chemical and organic characteristics of PM2.5 particles and their in-vitro cytotoxic effects on lung cells: the Middle East Dust storms in Ahvaz, Iran. Sci Total Environ. 2019;655:434-45. doi: 10.1016/j.scitotenv.2018.11.153.
- Deng XL, Shi CE, Wu BW, Yang YJ, Jin Q, Wang HL, et al. Characteristics of the water-soluble components of aerosol particles in Hefei, China. J Environ Sci (China). 2016;42:32-40. doi: 10.1016/j.jes.2015.07.010.
- Hien PD, Bac VT, Thinh NT. PMF receptor modelling of fine and coarse PM10 in air masses governing monsoon conditions in Hanoi, northern Vietnam. Atmos Environ. 2004;38(2):189-201. doi: 10.1016/j.atmosenv.2003.09.064.
- Sharma SK, Mandal TK, Jain S, Saraswati, Sharma A, Saxena M. Source apportionment of PM2.5 in Delhi, India using PMF model. Bull Environ Contam Toxicol. 2016;97(2):286-93. doi: 10.1007/s00128-016-1836-1.
- Khan MF, Shirasuna Y, Hirano K, Masunaga S. Characterization of PM2.5, PM2.5–10 and PM>10 in ambient air, Yokohama, Japan. Atmos Res. 2010;96(1):159-72. doi: 10.1016/j.atmosres.2009.12.009.
- Remoundaki E, Kassomenos P, Mantas E, Mihalopoulos N, Tsezos M. Composition and mass closure of PM2.5 in urban environment (Athens, Greece). Aerosol Air Qual Res. 2013;13(1):72-82. doi: 10.4209/aaqr.2012.03.0054.
- 43. Zhao PS, Zhang XL, Meng W, Yang BY, Fan WY, Liu HY. Characteristics of inorganic water-soluble ions from aerosols in Beijing-Tianjin-Hebei area. Environ Sci. 2011;32(6):1546-9.
- 44. Saraga D, Maggos T, Sadoun E, Fthenou E, Hassan H, Tsiouri V, et al. Chemical characterization of indoor and outdoor particulate matter (PM2.5, PM10) in Doha, Qatar. Aerosol Air Qual Res. 2017;17(5):1156-68. doi: 10.4209/ aaqr.2016.05.0198.
- 45. Dahmardeh Behrooz R, Esmaili-Sari A, Bahramifar N, Kaskaoutis DG, Saeb K, Rajaei F. Trace-element concentrations and water-soluble ions in size-segregated dust-borne and soil samples in Sistan, southeast Iran. Aeolian Res. 2017;25:87-105. doi: 10.1016/j. aeolia.2017.04.001.
- Zarasvandi A, Carranza EJM, Moore F, Rastmanesh F. Spatio-temporal occurrences and mineralogical-geochemical characteristics of airborne dusts in Khuzestan province (southwestern Iran). J Geochem Explor. 2011;111(3):138-51. doi: 10.1016/j.gexplo.2011.04.004.
- Gholampour A, Nabizadeh R, Hassanvand MS, Nazmara S, Mahvi AH. Elemental composition of particulate matters around Urmia Lake, Iran. Toxicol Environ Chem. 2017;99(1):17-31. doi: 10.1080/02772248.2016.1166226.
- Ho KF, Lee SC, Chan CK, Yu JC, Chow JC, Yao XH. Characterization of chemical species in PM2.5 and PM10 aerosols in Hong Kong. Atmos Environ. 2003;37(1):31-9. doi: 10.1016/s1352-2310(02)00804-x.
- Ho KF, Lee SC, Cao JJ, Chow JC, Watson JG, Chan CK. Seasonal variations and mass closure analysis of particulate matter in Hong Kong. Sci Total Environ. 2006;355(1-3):276-87. doi: 10.1016/j.scitotenv.2005.03.013.
- 50. Yan Y, Fu P, Jing B, Peng C, Boreddy SK, Yang F, et al. Hygroscopic behavior of water-soluble matter in marine aerosols over the East China Sea. Sci Total Environ. 2017;578:307-16. doi: 10.1016/j.scitotenv.2016.10.149.

- Hsu SC, Liu SC, Kao SJ, Jeng WL, Huang YT, Tseng CM, et al. Water-soluble species in the marine aerosol from the northern South China Sea: high chloride depletion related to air pollution. J Geophys Res Atmos. 2007;112(D19). doi: 10.1029/2007jd008844.
- 52. Samara C, Kantiranis N, Kollias P, Planou S, Kouras A, Besis A, et al. Spatial and seasonal variations of the chemical, mineralogical and morphological features of quasi-ultrafine particles (PM 0.49) at urban sites. Sci Total Environ. 2016;553:392-403. doi: 10.1016/j.scitotenv.2016.02.080.
- Szigeti T, Óvári M, Dunster C, Kelly FJ, Lucarelli F, Záray G. Changes in chemical composition and oxidative potential of urban PM2.5 between 2010 and 2013 in Hungary. Sci Total Environ. 2015;518-519:534-44. doi: 10.1016/j. scitotenv.2015.03.025.
- 54. Guo YT, Zhang J, Wang SG, She F, Li X. Long-term characterization of major water-soluble inorganic ions in PM10 in coastal site on the Japan Sea. J Atmos Chem. 2011;68(4):299-316. doi: 10.1007/s10874-012-9223-8.
- 55. Rajput P, Sarin M, Sharma D, Singh D. Characteristics and emission budget of carbonaceous species from postharvest agricultural-waste burning in source region of the Indo-Gangetic Plain. Tellus B Chem Phys Meteorol. 2014;66(1):21026. doi: 10.3402/tellusb.v66.21026.
- 56. Wang Y, Zhuang G, Tang A, Yuan H, Sun Y, Chen S, et al. The ion chemistry and the source of PM2.5 aerosol in Beijing. Atmos Environ. 2005;39(21):3771-84. doi: 10.1016/j.atmosenv.2005.03.013.
- Zhou J, Xing Z, Deng J, Du K. Characterizing and sourcing ambient PM2.5 over key emission regions in China I: watersoluble ions and carbonaceous fractions. Atmos Environ. 2016;135:20-30. doi: 10.1016/j.atmosenv.2016.03.054.
- Yao X, Chan CK, Fang M, Cadle S, Chan T, Mulawa P, et al. The water-soluble ionic composition of PM2.5 in Shanghai and Beijing, China. Atmos Environ. 2002;36(26):4223-34. doi: 10.1016/s1352-2310(02)00342-4.
- 59. He Q, Yan Y, Guo L, Zhang Y, Zhang G, Wang X. Characterization and source analysis of water-soluble inorganic ionic species in PM2.5 in Taiyuan city, China. Atmos Res. 2017;184:48-55. doi: 10.1016/j. atmosres.2016.10.008.
- Kang CH, Kim WH, Ko HJ, Hong SB. Asian Dust effects on total suspended particulate (TSP) compositions at Gosan in Jeju Island, Korea. Atmos Res. 2009;94(2):345-55. doi: 10.1016/j.atmosres.2009.06.013.
- Wang Y, Zhuang G, Zhang X, Huang K, Xu C, Tang A, et al. The ion chemistry, seasonal cycle, and sources of PM2.5 and TSP aerosol in Shanghai. Atmos Environ. 2006;40(16):2935-52. doi: 10.1016/j.atmosenv.2005.12.051.
- Deshmukh DK, Deb MK, Tsai YI, Mkoma SL. Water soluble ions in PM2.5 and PM1 aerosols in Durg city, Chhattisgarh, India. Aerosol Air Qual Res. 2011;11(6):696-708. doi: 10.4209/aaqr.2011.03.0023.
- 63. Shon ZH, Kim KH, Song SK, Jung K, Kim NJ, Lee JB. Relationship between water-soluble ions in PM2.5 and their precursor gases in Seoul megacity. Atmos Environ. 2012;59:540-50. doi: 10.1016/j.atmosenv.2012.04.033.
- 64. Manousakas M, Papaefthymiou H, Diapouli E, Migliori A, Karydas AG, Bogdanovic-Radovic I, et al. Assessment of PM2.5 sources and their corresponding level of uncertainty in a coastal urban area using EPA PMF 5.0 enhanced diagnostics. Sci Total Environ. 2017;574:155-64. doi: 10.1016/j.scitotenv.2016.09.047.

- Tepe AM, Doğan G. Chemical characterization of PM2.5 and PM2.5–10 samples collected in urban site in Mediterranean coast of Turkey. Atmos Pollut Res. 2021;12(1):46-59. doi: 10.1016/j.apr.2020.08.012.
- 66. Li R, Wang Q, He X, Zhu S, Zhang K, Duan Y, et al. Source apportionment of PM2.5 in Shanghai based on hourly organic molecular markers and other source tracers. Atmos Chem Phys. 2020;20(20):12047-61. doi: 10.5194/ acp-20-12047-2020.
- Wang M, Li X, He WY, Li JX, Zhu YY, Liao YL, et al. Distribution, health risk assessment, and anthropogenic sources of fluoride in farmland soils in phosphate industrial area, southwest China. Environ Pollut. 2019;249:423-33. doi: 10.1016/j.envpol.2019.03.044.
- Qurat-ul-Ain, Farooqi A, Sultana J, Masood N. Arsenic and fluoride co-contamination in shallow aquifers from agricultural suburbs and an industrial area of Punjab, Pakistan: spatial trends, sources and human health implications. Toxicol Ind Health. 2017;33(8):655-72. doi: 10.1177/0748233717706802.
- 69. Hao Y, Gao C, Deng S, Yuan M, Song W, Lu Z, et al. Chemical characterisation of PM2.5 emitted from motor vehicles powered by diesel, gasoline, natural gas and methanol fuel. Sci Total Environ. 2019;674:128-39. doi:

10.1016/j.scitotenv.2019.03.410.

- 70. Cui M, Chen Y, Feng Y, Li C, Zheng J, Tian C, et al. Measurement of PM and its chemical composition in real-world emissions from non-road and on-road diesel vehicles. Atmos Chem Phys. 2017;17(11):6779-95. doi: 10.5194/acp-17-6779-2017.
- Saraga DE, Tolis EI, Maggos T, Vasilakos C, Bartzis JG. PM2.5 source apportionment for the port city of Thessaloniki, Greece. Sci Total Environ. 2019;650(Pt 2):2337-54. doi: 10.1016/j.scitotenv.2018.09.250.
- 72. Kikuchi R, Takada M, Hifumi K, Yoshimura K, Ozeki T, Kimoto T, et al. The degree of Cl-loss for the particulate matter (PM) and fog water sampled at the same air mass at the Hachimantai mountain range in northern Japan. Atmos Res. 2009;94(3):501-9. doi: 10.1016/j.atmosres.2009.08.001.
- 73. Zhao Y, Gao Y. Acidic species and chloride depletion in coarse aerosol particles in the US east coast. Sci Total Environ. 2008;407(1):541-7. doi: 10.1016/j. scitotenv.2008.09.002.
- 74. Contini D, Genga A, Cesari D, Siciliano M, Donateo A, Bove MC, et al. Characterisation and source apportionment of PM10 in an urban background site in Lecce. Atmos Res. 2010;95(1):40-54. doi: 10.1016/j.atmosres.2009.07.010.