

Comparison of insect biological diversity parameters in natural and degraded habitats in Gölcük Nature Park (Isparta, Turkey)

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Abstract: The study was carried out to compare the biodiversity parameters of the insect species in natural and un-natural habitats in Gölcük Nature Park (Isparta Province) between 2018 and 2019. Two natural and two unnatural habitats (impact of tourism) were selected to realize mentioned aim. Totally 40 pitfall traps (10 traps for each habitat) were set up to sample Coleoptera species. Species richness was found higher in natural habitats than unnatural ones. Diversity index (both Shannon-Wiener and Simpson) were calculated higher in natural habitats and as expected, dominancy inversely proportional to diversity was measured lower in natural habitats. Shannon Evenness, population density due to individual of the species showed mostly evenly distributed on the graphs in natural habitats. The Sörenson Coefficient results showed that selected natural and unnatural habitats groups were found similar to each other. The results of the study showed that human activity plays an important role in habitat destruction. It has been revealed that species richness and diversity are negatively affected in habitats with human impact.

Keywords: Coleoptera, Shannon-Wiener, Simpson, Sörenson, Evenness

Gölcük Tabiat Parkın'da doğal ve bozulmuş habitatlarında böcek biyolojik çeşitlilik parametrelerinin karşılaştırılması (Isparta, Türkiye)

Özet: Calısma, 2018-2019 yılları arasında Gölcük Tabiat Parkı'nda (Isparta İli) doğal ve doğal olmavan habitatlarda bulunan böcek türlerinin biyoçeşitlilik parametrelerinin karşılaştırılması amacıyla yapılmıştır. Bahsedilen amaca ulaşmak için iki doğal ve iki doğal olmayan (turizmin etkisi) orman ekosistemi seçilmiştir. Coleoptera türlerini örneklemek için toplam 40 adet çukur tuzak (her habitat için 10'ar adet) kurulmuştur. Çalışma sonunda doğal habitatlarda tür zenginliği doğal olmayanlara göre daha yüksek bulunmustur. Çeşitlilik indeksi (hem Shannon-Wiener hem de Simpson) doğal habitatlarda daha yüksek hesaplanmış ve beklendiği gibi, çeşitlilikle ters orantılı olan baskınlık, doğal olmayan habitatlarda daha yüksek ölçülmüştür. Shannon Evenness, popülasyon yoğunluk ilişkisi doğal yaşam ortamlarında çoğunlukla eşit dağılımlı grafikler oluşturmuştur. Sörenson katsayısı sonuçları, seçilen doğal ve doğal olmayan ekosistemlerin gruplar halinde birbirine benzer olduğunu göstermiştir. Çalışmanın sonuçları, insan faaliyetinin habitat tahribatında önemli bir rol oynadığını göstermiştir. İnsan etkisi ile habitatlarda tür zenginliği ve çeşitliliğinin olumsuz etkilendiği ortaya konmuştur.

Anahtar kelimeler: Coleoptera, Shannon-Wiener, Simpson, Sörenson, Evenness

1. Introduction

Turkey was formed by the coming together of the different geographical zones thus different habitats have enabled different animal and plant species to live there. That's why Turkey is the one of the most important areas in planet in terms of species richness as a natural consequence of this diversity (Aydin, 2006).

Gölcük Nature Park is the one of the six protected area in Isparta Province in Turkey. Gölcük Nature Park is located in Lakes Areas, Turkey's most important biodiversity areas, in the western part of Taurus. It is located on the transition zone between the Mediterranean and the terrestrial climate, and its altitude varies between 900-2000 meters. Extensive studies have been carried out in the Gölcük Nature Park in terms of plant diversity, but unfortunately almost no scientific study has been done in terms of insect diversity in this valuable area. Studies have shown that there are 227 species of plants

belonging to 47 families in the mentioned protected area (Fakir and Dutkuner, 1999). However, the biodiversity values in these areas are negatively affected due to the richness of species that have recently been destroyed by human activity.

In this study, the habitats of human activity were compared with natural similar ones, and the reasons for the difference between natural and destructed habitats were discussed by calculating the biological diversity parameters.

2. Materials and methods

2.1. Study areas

This study was carried out in Gölcük Nature Park from May to September between 2018 and 2019. The habitats dominated by with Pinus nigra Arn. subsp. pallasiana (Lamb.) Holmboe were chosen as study area. Two natural

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- Received (Geliş tarihi): 07.10.2021, Accepted (Kabul tarihi): 20.11.2021



Citation (Atıf): Aydın, G., 2021. Comparison of insect biological diversity parameters in natural and degraded habitats in Gölcük Nature Park (Isparta, Turkey). Turkish Journal of Forestry, 22(4): 362-365. DOI: 10.18182/tjf.1006039

(NF_1, NF_2) and two unnatural habitat (effected of tourism) (UnNF_1, UnNF_2) were selected to compare differentiation of habitats.

2.2. Sampling method

Totally 40 pitfall traps (10 traps for each habitat) are placed in all habitat. Plastic containers of approximately 15 cm in diameter and 20 cm in depth were buried in the soil in habitats, with 10 pieces at 10 meters intervals, with the open parts at the soil level (New, 1998). Biodiversity parameters were calculated with the data obtained from pitfall traps. Sampled insects that fell into pitfall traps were killed with the help of a killing bottle and brought to the "Insect Biodiversity Laboratory" in Isparta University of Applied Sciences, Atabey Vocational School for labelling, pinning, and counting. Distinctions were made on the basis of family, genus and species level.

2.3. Measurement of biological diversity parametres

Biodiversity basic parameters of habitats were calculated using the EvenDiv 1.1 program (Heimann, 2004), the parameters used and their calculation methods are given below:

Shannon-Wiener and Simpson diversity indices were used to determine species diversity.

- Shannon-Wiener diversity index (H')

$$H' = -\sum p_i \ln(p_i) \tag{1}$$

where H' is the index of diversity, pi is the importance value of a species as a proportion of all species, and ln is the natural logarithm

- Simpson diversity index (S)

$$S = 1 - \sum n_i(n_i - 1) / N(N - 1)$$
⁽²⁾

where S is the index of diversity, ni is the importance value of a species as a proportion of all species, and N is the sum of the number of individuals (Magurran, 1988; Magurran, 2004; Özkan, 2016).

Simpson dominancy index was used to determine dominance

- Simpson's dominance index (Sd)

$$Sd = \sum n_i(n_i - 1) / N(N - 1)$$
 (3)

where Sd is the index of dominancy, i is number of species, ni is the importance value of a species as a proportion of all species, and N is the sum of the number of individuals

Shannon Evenness and Simpson Evenness index were used to determine population density relationships of the species.

- Shannon evenness index (EH)

$$EH = H'/\ln(N)) \tag{4}$$

where EH is Evenness index, H' is the index of Shannon-Wiener diversity, In is the natural logarithm, and N is the sum of the number of individuals.

- Simpson Evenness index (Esm)

$$Esm = S / N$$
(5)

where Esm is Simpson Evenness, S is Simpson diversity, N is species richness (Magurran, 1988 ve 2004; Özkan, 2016).

Percentage similarity index (Bs) was used to determine the compositional similarity between the habitats

- Percent similarity

$$\% S = \sum \min(a, b, \dots, x) \tag{6}$$

where %S is percent similarity, \sum min is the sum of the smallest values whose percentages are calculated in the habitat with the smallest values in the other habitat whose similarity is calculated (Kreps, 1999).

Multi Variate Statistical Package (MVSP) 3.11c program was used to classify selected habitats (Kovach, 1999).

The identification of the insects sampled at the family level is carried out by Borror et al. (1981). Insects diagnosed at the family level were identified at the "morpho-species" level (Lodge and Cantrell, 1995; Clauson, 2002, Ryder et al., 2005; Borgelt and New, 2006; Dudgeon, 2006; Yanoviak et al., 2006; Grimbacher and Stork, 2007).

3. Results

During the study, 1544 individuals belonging to 50 species from 13 different families under Coleoptera were sampled (see appendix). Species richness were found much higher in natural habitats than habitat under tourism activity in both study years (Table 1). The most individuals were collected in NF_2 with 218 and 264 in 2018 and 2019, respectively. Shannon-Wiener were found the highest values in natural habitats in both NF_1 and NF_2 with 3.39 and 3.46 in 2018 and 3.49 and 3.52 in 2019. Simpson diversity index gave similar results around 0.96 ± 1 in both study years (Table 1).

Table 1. Biodiversity parameter values calculated with the data obtained from the pitfall trap sampling method in Natural *Pinus nigra* Forest (NF_1 and NF_2) and un-natural *Pinus nigra* Forest effected by tourism activity (UnNF_1 and UnNF_2) between 2018 and 2019.

	2018						
	NF_1	NF_2	UnNF_1	UnNF_2			
No. of Species	39	41	28	28			
No. of Individuals	185	218	130	141			
Diversity indices							
Shannon-Wiener[H]	3.3956	3.4602	2.9276	2.7147			
Simpson Index[D]	0.0426	0.0393	0.0755	0.1065			
Simpson Diversity[1-D]	0.9574	0.9607	0.9245	0.8935			
Evenness indices							
Shannon-Evenness[EH]	0.9269	0.9318	0.8786	0.8147			
Simpson-Evenness [E1/D]	0.6019	0.6206	0.473	0.3353			
	2019						
No. of Species	43	43	29	30			
No. of Individuals	206	264	187	213			
Diversity indices							
Shannon-Wiener[H]	3.4992	3.5296	3.0025	2.8112			
Simpson Index[D]	0.0415	0.035	0.0715	0.0991			
Simpson Diversity[1-D]	0.9585	0.9650	0.9285	0.9009			
Evenness indices							
Shannon-Evenness[EH]	0.9303	0.9384	0.8917	0.8265			
Simpson-Evenness [E1/D]	0.5604	0.6645	0.4823	0.3364			

In contrast to diversity, dominance values were found higher in habitats with tourism activity. Result of the Evenness index, both Shannon and Simpson, showed that individual of the species living in natural habitats were more balanced distribution than un-natural ones although obtained values were found close to each other (Table 1).

Similarity index found that naturally selected habitats (NF_1 and NF_2) have higher similarities to each other than unnatural habitats (UnNF_1 and UnNF_2) during the first study year while un-naturel habitats were found more similar than natural ones in 2019 (Figure 1).

4. Discussion

The reason why the species richness and other biodiversity values measured in habitats that have been exposed to human activity for a long time can be measured lower when compared to natural habitats can be explained the displacement of the species and the increase and decrease of their populations (Aydin et al., 2005; Aydin and Kazak, 2007 ve 2010; Aydin and Karaca, 2011; Aydin, 2018). In this study, although the biodiversity measurement results show high biodiversity values in natural habitats, the biodiversity and evenness calculated values in unnatural habitats are found to be close to each other, which cannot be underestimated.

The present study showed that the habitat changes due to tourism activities had changed not only to the insect diversity but also to their structure within communities (see appendix). Many scientific studies showed similar results (Morris, 2010; Sutrisno, 2010; Barron et al., 2019; Forister et al., 2019).

Similarity results showed different results in both years. This is because the study was conducted for only two years. Long-term biological diversity measurements in the same habitats are necessary to obtain results that can explain the reason for the difference.



Figure 1. Percentage similarity classification analysis calculated considering the insect species sampled in Natural *Pinus nigra* Forest (NF_1 and NF_2) and un-natural *Pinus nigra* Forest effected by tourism activity (UnNF_1 and UnNF_2) (UPGMA: Unweighted Pair Group with Arithmetic Mean) * 2018 upper; 2019 bottom.

5. Conclusion

Biodiversity parameter values to be measured each year in order to ensure the sustainability of protected areas may indicate whether the habitats are degraded or not. Even though insects are not used frequently in the sustainability of protected areas, benefit can be gained by selecting insect groups that sustain their lives on the soil surface such as Carabidae, Tenebrionidae, Scarabaeidae, Staphylinidae, etc. and have high potential to be used as biological indicators.

The use of insect species in habitat protection can be achieved by using them as a single indicator, as well as by considering insect assemblages as a whole. This can be achieved by periodically measuring biological diversity parameters in order to carry protected areas into the future.

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Appendix. The list of Coleoptera species sampled by pitfall traps during 2018 and 2019 in Gölcük Nature Park in Isparta, Turkey.

Family	Species	NF_1	NF_2	UnNF_1	UnNF_2
Apionidae	Apion sp. Herbst, 1797	10	6	0	0
Buprestidae	Anthaxia diadema (Fischer von Waldheim, 1824)	0	1	3	3
*	Capnodis miliaris (Klug, 1829)	3	0	6	10
	Julodis ehrenbergii Laporte, 1835	4	9	27	10
	Trachypteris picta (Pallas, 1773)	3	5	7	9
Carabidae	Abax sp. Bonelli, 1810	16	22	0	0
	Abax sp. Bonelli, 1810	2	0	0	0
	Amara sp. Bonelli, 1810	13	11	0	0
	Amara sp. Bonelli, 1810	6	3	47	71
	Amara sp. Bonelli, 1810	2	1	0	1
	Acinopus sp. Dejean, 1821	12	27	5	10
	Agonum sp. Bonelli, 1810	5	8	5	7
	Carabus cribratus (Motschulsky, 1850)	9	19	7	3
	Carabus glabratus Paykull, 1790	50	39	0	0
	Carabus graecus Dejean, 1826	20	39	11	5
	Carabus gotschii Chaudoir, 1846	7	11	0	0
	Carabus scabripennis Chaudoir, 1850	12	20	0	0
	Carabus tenuitarsis (Kraatz, 1877)	3	8	3	5
	Carabus victor Fischer von Waldheim, 1836	14	15	4	5
	Harpalinus sp. Jeannel, 1946	7	17	4	7
	Lamprostus torosus (I.Frivaldszky von Frivald, 1835)	12	11	0	0
	Pachystus sp. Motschoulsky, 1865	9	17	0	0
	Pachystus graecus (Dejean, 1826)	8	10	7	6
	Procrustes anatolicus Chaudoir, 1857	13	13	0	0
	Procerus scabrosus (A.G.Olivier, 1790)	5	7	11	3
Cerambycidae	Cerambyx sp. Linnaeus, 1758	0	4	8	7
	Stictoleptura excisipes (K.Daniel & J.Daniel, 1891)	0	4	5	10
Cetoniidae	Cetonia aurata (Linnaeus, 1758)	11	14	18	11
Chrysomelidae	Chrysolina herbacea (Duftschmid, 1825)	9	14	11	3
	Chrysomela populi Linnaeus, 1758	3	0	0	4
	Cryptocephalus duplicatus Suffrian, 1845	0	3	0	0
	Cryptocephalus flavipes Fabricius, 1781	1	0	1	0
	Labidostomis sp. Germar, 1817	7	3	0	0
Cicindelidae	Cicindela campestris Linnaeus, 1758	7	8	0	0
Coccinellidae	Adalia bipunctata (Linnaeus, 1758)	7	5	0	0
	Coccinella septempunctata Linnaeus, 1758	15	7	15	47
	Exochomus quadripustulatus (Linnaeus, 1758)	7	9	11	5
Curculionidae	Larinus onopordi (Fabricius & J.C., 1787)	6	0	0	0
	Otiorhynchus sp. Germar, 1822	9	13	7	16
	Phyllobius incanus Gyllenhal, 1834	5	11	0	0
	Sitona macularius (Marsham, 1802)	0	0	5	9
Melolonthidae	Anoxia asiatica Desbrochers, 1871	14	18	14	5
	Melolontha melolontha (Linnaeus, 1758)	8	4	9	7
	Polyphylla fullo (Linnaeus, 1758)	0	7	4	2
Scarabaeidae	Scarabaeus sacer Linnaeus, 1758	4	8	50	64
Scolytidae	Blastophagus minor (Hartig, 1834)	7	0	0	0
	Blastophagus piniperda Eichhoff, 1864b	1	5	5	3
	Ips sp. De Geer, 1775	5	10	7	6
Tenebrionidae	Colpotus vogti Koch, 1944	0	7	0	0
	Tenebrio sp. Linnaeus, 1758	20	9	0	0