

Rice straw open burning: emissions, effects and multiple benefits of non-burning alternatives

Nguyen Thi Kim Oanh*

*Environmental Engineering and Management, School of Environment, Resources and Development, Asian Institute of Technology, Pathumthani, Thailand
Asian Institute of Technology Center in Vietnam, Vietnam*

Received 22 January 2021; accepted 5 April 2021

Abstract:

Rice is one of the most important staple foods not just to people in Asia, but around the world. To meet domestic and export demands, farmers in Southeast Asia (SEA) grow 2-3 crop cycles per year, which leaves only a short period for land preparation. Field open burning of rice straw has been widely practiced to quickly clear the surface biomass for the next crop planting. However, this uncontrolled open combustion of rice straw releases large amounts of toxic air pollutants including key conventional pollutants along with carcinogenic compounds like dioxins, polycyclic aromatic hydrocarbons, and benzene, as well as major climate forcing agents. Emissions from rice straw open burning (RSOB) have been shown to significantly elevate ambient levels of $PM_{2.5}$ and surface ozone in adjacent urban areas. During the dry season, when stagnant meteorological conditions are prevalent, intensive open burning activities further intensify haze episodes. Rice straw, however, is a valuable resource that should be recovered and not disposed of by open burning. Indeed, several non-open burning alternatives are available that would bring in multiple benefits to air quality, climate, health, and economy. For example, the production of rice straw fuel pellets for cooking in clean gasifier cookstoves is one promising option. For the successful elimination of RSOB in SEA, technology development along with formulation and implementation of appropriate policies should be in place to mobilise active participation from all stakeholders.

Keywords: co-benefits, cooking, rice straw open burning, rice straw pellet, toxic pollutants.

Classification number: 5.3

Introduction

Rice is the most popular crop in Asia with an annual production comprising of 90% of the world's total production. Rice is also the most important staple food in Asia as it provides 50-80% of the total calories consumed. Several countries in SEA are among the world's top ten rice exporters and, presently, the region collectively produces over 200 million metric tons (tonnes) of rice annually. To meet domestic and export demands, SEA farmers grow 2-3 crop cycles per year. Hence, there is only a short time period left to prepare the land for planting the next crop. Post-harvest RSOB has long been widely utilized by farmers in the region as it can quickly clear the surface biomass to facilitate land preparation. Survey results have shown that farmers in SEA prefer RSOB for land preparation because it requires less labor and also helps control undesirable weeds and pests along with providing ash as nutrients back into the soil [1]. Meanwhile, as farmers become wealthier and farming work becomes more mechanized, the demand for crop residue as cooking fuel or animal feedstock is declining. As a result, RSOB activity is now widespread and during the harvesting period the effects of smoke are felt on both local and regional scales, especially in the dry season.

By nature, RSOB is an uncontrolled combustion of vegetation biomass at low temperatures, hence, huge amounts of products of incomplete combustion (PIC) are released. These PIC are toxic air pollutants and include, for example, particulate matter (PM) that are mainly composed of fine inhalable particles or $PM_{2.5}$ (particles with aerodynamic diameters $\leq 2.5 \mu m$) together with black carbon (BC) and organic carbon (OC) components (among others), gaseous pollutants of carbon monoxide (CO), and a range of volatile organic compounds (VOCs). A wide range of toxic and carcinogenic semi-volatile organic compounds (SVOCs that present in both PM and gas phases) such as polychlorinated dibenzo-*p*-dioxins and dibenzofurans (PCDD/PCDFs, hereafter referred to as dioxins), polycyclic aromatic hydrocarbons (PAHs), and organochlorinated pesticides are also found in RSOB smoke [2-13]. Many of these SVOCs are persistent organic pollutants that present over an extended time in the environment and have the ability to bioaccumulate in tissues, which makes them even of more of a health concern [14, 15].

It is worth mentioning that important greenhouse gases (GHGs) like methane (CH_4) and nitrous oxide (N_2O) are also

*Email: kimoanh@ait.asia

emitted from RSOB. While a significant amount of CO_2 is also released from the activity, the cycle is considered climate neutral because it is absorbed by the growth of the next crop. Besides, several toxic pollutants emitted from RSOB, so called “short-lived climate forcers” or “short-lived climate pollutants” (SLCPs), also have climate forcing effects. For example, BC is a strong climate warming agent while OC is a cooling agent [16]. In addition, pollutants released from RSOB can participate in chemical reactions in the atmosphere to form other pollutants that have both air quality and climate effects. As an example, VOCs and NO_x in the presence of sunlight participate in photochemical reactions to form the tropospheric ozone, a secondary pollutant that is not only a strong GHG but also a toxic air pollutant to human health and plants and hence can reduce crop yield [17]. Nitrogen oxides ($\text{NO}_x = \text{NO} + \text{NO}_2$) and sulfur oxides (SO_x) released from the activity are important precursors of secondary inorganic particles, while VOCs are precursors of secondary organic particles. These secondary particles are formed in the atmosphere and they belong to the $\text{PM}_{2.5}$ size range.

There are two common burning methods for rice straw currently observed in SEA, namely pile burning and spread burning, which have different emission amounts per kg of rice straw burned [18, 19]. Pile burning is typically practiced after manual harvesting when RS is piled up at a paddy corner (or sometimes inside villages) and burned largely under smoldering conditions with a visible dense smoke plume, containing huge amounts of toxic pollutants (Fig. 1A). Spread burning (Fig. 1B) is normally applied in places where mechanical harvesting equipment is used. The combine harvesters cut the upper parts of rice plants and spread them in windrows while leaving the lower parts (or standing parts) virtually untouched. Spread burning fires normally consume most of the spreading RS but the standing parts, in many cases, are only partially burned especially when RS moisture is high. In Vietnam, pile burning is commonly practiced in the Red river delta region. Emissions from RSOB also strongly depend on the combustion conditions, which, in turn, depend on the moisture content of the RS and paddy soil (e.g., the higher moisture the more emissions), winds, and air humidity, among others.

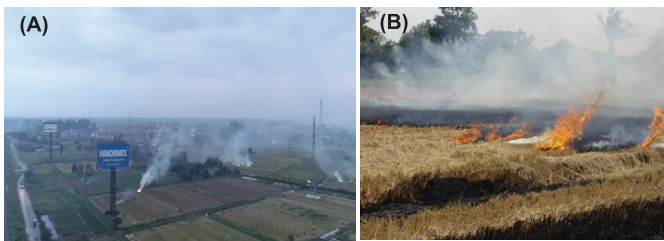


Fig. 1. Rice straw field burning: (A) pile burning under smoldering conditions around Hanoi, Vietnam, May 2020 (Source: Dan Tri newspaper) and (B) spread burning in Pathumthani, Thailand (photo by author).

Results of emission factors from experimental studies

The results of emission factor measurements from the spread RSOB experiments conducted in Thailand [6, 18] are presented in Table 1 for particulate, gaseous, and SVOC pollutants. Other references quoted in Table 1 provide emission factors for RSOB but without indicating the burning methods, namely the spread or pile burning. The emission factors for burning of agricultural crop residues in general, compiled from different laboratory studies by Andreae and Merlet (2001) [12], are also presented in Table 1.

Table 1. Emission factors of pollutants from RSOB (average \pm SD), mass of pollutants (g or mg as specified) per kg of dry rice straw.

| Pollutants | Rice straw | | General agro residue (f) |
|---------------------------------|---|--|---|
| | Spread RSOB [6, 18] | Other data sources | |
| Particulates | | | |
| PM _{2.5} , g/kg | 8.3±2.2 | 3.8 (a) | 3.9 |
| EC, g/kg | 0.53 | | 0.69±0.13 |
| OC, g/kg | 2.78 | | 3.3 |
| Water soluble ions*, g/kg | 1.5 | | |
| Levoglucosan, g/kg | 0.47 | | 0.27 |
| Total elements**, g/kg | 0.23 | | |
| Gaseous pollutants | | | |
| CO, g/kg | 93±10 | 180±40 (b); 64±5 (c) | 92±84 |
| CO ₂ , g/kg | 1177±140 | 1216±97 (b); 791±13 (c) | 1515±177 |
| Benzene, mg/kg | 763±266 | 870±200 (b) | 140 |
| Toluene, mg/kg | 232±3.4 | 1080±350 (b) | 26 |
| Ethylbenzene, mg/kg | nd | | 30 |
| Xylenes, mg/kg | nd | | 10 |
| SO ₂ , mg/kg | 510±320 | 180±310 (g) | 400 |
| NO _x , mg/kg | 490±210 (NO ₂) 1120±480 (NO _x) | 790±50 (c, NO ₂); 620±400 (b, NO) | 2500±100 (NO _x) |
| Aldehydes, mg/kg | 147±8.0 (hood burning) | 3170±880 (b, HCHO) | 1400 (HCHO) |
| Semi-volatile organic compounds | | | |
| PAHs (16 USEPA), mg/kg | Particulate | 34±35 | 1.02 (a); 18.6 (d); 3.0 (5% moist.) & 17.2 (20% moist.) (e) |
| | Gaseous | 230±333 | |
| | Total | 264±335 | 17.8 (a) |
| OCPs, mg/kg | Particulate | 0.086±0.052 | |
| | Gaseous | 0.141±0.194 | |
| | Total | 0.227±0.20 | |

Data sources: a: [13]; b: [20], values for dry fuel, ash free; c: [21]; d: [22]; e: [23] for 5% and 20% moisture content of RS, included 9 PPAHs (Fth, BaA, BbF, BfF, BkF, BaP, IcdP, DahA and BghiP); f: [12]: estimates based on laboratory studies; g: [24].

*: total 9 water soluble ions (sodium, potassium, ammonium, magnesium, calcium, fluoride, chloride, nitrate, and sulfate).

**: total of 33 detected elements (including four more elements, i.e. Ca, Mg, Cl and K, than that presented in N.T. Kim Oanh, et al. (2011) [18]).

There are very few experimental studies reporting emission factors of pile RSOB. N.T. Kim Oanh, et al. (2011) [18] reported the emission factor of $PM_{2.5}$ for pile RSOB, which was about 18.3 g/kg of dry rice straw, i.e., more than two times greater than that of spread burning (8.3 g/kg). Based on the PM source profile reported in N.T. Kim Oanh, et al. (2011) [18], the estimated emission factors of elemental carbon (EC, an operational definition of BC) is about 1.1 g/kg while that of OC is about 6.1 g/kg.

Emissions from RSOB and effects on local air quality

Emissions from RSOB in SEA and Vietnam: during 2010-2015, about 120 million tonnes of rice straw in SEA was disposed of annually by open burning with about 24 million tonnes from Vietnam alone [5]. Huge emissions are released from RSOB activity in SEA; roughly 1.8 million tonnes of $PM_{2.5}$, 12 million tonnes of CO, 65 g I-TEQ of dioxins (in the unit of toxicity equivalent, to the most toxic congener 2, 3, 7, 8-TCDD, based on the international toxicity scale), 25 thousand tonnes of PAHs, and 29 tonnes of OCPs, together with other toxic air pollutants such as benzene, toluene, and aldehydes (Table 2). By country, in the descending order, Indonesia, Vietnam, Myanmar, Thailand, and Philippines had the largest shares in RSOB emissions and collectively contributed more than 95% of the total SEA emissions from this activity. The emissions from RSOB in Vietnam typically contribute about 16-20% of the SEA's total and varies with pollutants.

N.T. Kim Oanh, et al. (2018) [5] considered crop residue open burning (CROB) of the eight main crop types in SEA and reported that RSOB was the major contributor to CROB emissions by sharing 70-95% of the total amounts of different pollutants released. In 2010, CROB emissions in SEA contributed less than forest fires to the total emissions from these two major biomass open burning source categories (SUM=CROB+forest fires), i.e., CROB contributed about 10-43% to the SUM, varies with species. However, the shares of CROB emissions in the SUM differ significantly between countries. In Vietnam, for example, emissions from CROB were generally higher than the forest fires, i.e., sharing 49-92% of the SUM. In the Philippines, contributions from those two sources were in similar ranges, with the shares ranging between 33-69% that varies with species. However, forest fires had much higher contributions to the SUM in the countries of Indonesia, Thailand, and Myanmar [5]. It is worth emphasising that the effects of CROB emissions receive much less attention from society than catastrophic SEA transboundary haze events caused by forest fires.

Table 2. Annual emissions (in specified units) from RSOB in SEA and Vietnam averaged over the period 2010-2015.

| Pollutants | Unit (*) | Vietnam | SEA |
|-------------------|------------|---------|--------|
| CO | Gg/yr | 2258 | 12000 |
| NO _x | Gg/yr | 55 | 290 |
| SO ₂ | Gg/yr | 4.4 | 23 |
| NM VOC | Gg/yr | 170 | 890 |
| NH ₃ | Gg/yr | 100 | 500 |
| PM ₁₀ | Gg/yr | 330 | 2000 |
| PM _{2.5} | Gg/yr | 295 | 1800 |
| BC | Gg/yr | 13.6 | 70 |
| OC | Gg/yr | 228 | 800 |
| CO ₂ | Gg/yr | 28600 | 153900 |
| CH ₄ | Gg/yr | 100 | 520 |
| N ₂ O | Gg/yr | 2.4 | 13 |
| Aldehydes | Gg/yr | 3.6 | 19 |
| Benzene | Gg/yr | 18.5 | 97 |
| Toluene | Gg/yr | 5.6 | 29 |
| PAHs | Gg/yr | 6.4 | 25 |
| OCPs | t/yr | 5.5 | 29 |
| Dioxins | g I-TEQ/yr | 13 | 65 |

*Gg: thousand tonnes. Source: adapted from N.T. Kim Oanh, et al. (2018) [5].

Effects of RSOB on ambient air quality: several studies in SEA show evidence of the effects of RSOB emissions on local air quality. In Pathumthani, a large rice growing province of Thailand, RSOB is especially intensive during the dry months from November to April. The levels of carcinogenic PAHs measured in the ambient air of a rural area of Klong Luang (KL) district during the intensive RSOB days were above 400 ng/m³ that is 60 times higher than the levels measured in the air in a remote site of Khao Yai (KY) national park [4]. The ambient profile of PAHs in KL during RSOB days (Fig. 2) shows a dominance of 4-ring compounds, most remarkably

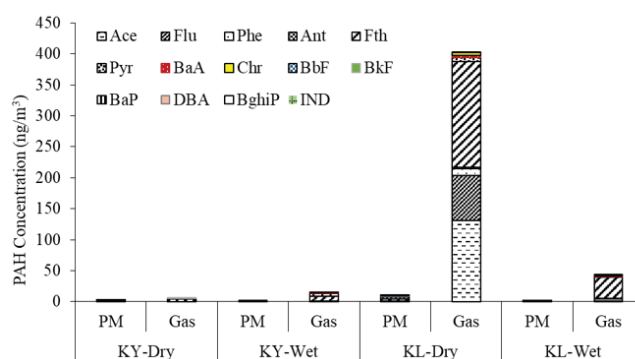


Fig. 2. PAHs levels and individual compound profiles in an intensive RSOB area (KL) and the remote national park (KY) during the dry and wet season (measurement data, extracted from Tipayarom and N.T. Kim Oanh (2020) [4]).

fluoranthene, which indicates a strong influence of rice straw burning emissions on the PAH air quality [3, 25, 26]. The RSOB smoke in KL also contains high levels of OCPs (about 14 ng/m³), which may be related to the re-emission of compounds accumulated in the paddy soil from the past applications.

Based on the analysis of mass and compositions of PM_{2.5} measured in KL using the receptor modeling approach, ambient PM_{2.5} levels show significant influence from RSOB. During the dry season, RSOB contributed about 14 µg/m³ to the PM_{2.5} mass concentration (40% of the total measured PM_{2.5} mass), which is well above that estimated in the wet season with 4 µg/m³ (25% of PM_{2.5} mass) [27]. The air quality dispersion modeling using a 3D chemical transport model (CAMx-MM5) also revealed impacts on the surface ozone air quality by RSOB in the Bangkok metropolitan region [28]. The simulation results for an ozone episode in March showed that RSOB in the modeling domain would cause an increase in the hourly ozone, by an average 4 ppb with a maximum of 10 ppb, in the Rangsit station (located near KL), over the scenario with zero RSOB.

Exposure to RSOB smoke and potential health effects: rice straw field burning emissions have been reported to induce high personal exposure to the toxic air pollutants in Asia, Europe, and the US [29-32]. However, the health effects specifically induced by RSOB have not been intensively studied. Torigoe, et al. (2000) [33] conducted a survey and revealed that emissions from RSOB in a study area in Japan possibly induced or exacerbated asthma attacks in children. The authors recommended the elimination of RSOB activity for the protection of the inhabitants' health, especially of children with asthma. Overall, intensive agricultural waste burning activities can affect air quality, human health, and the climate at continental and global scales [34, 35].

In SEA, large amounts of toxic air pollutants (PM_{2.5}, CO, VOCs, PAHs, OCPs, dioxins, etc.) are annually released from RSOB. The high levels of toxic air pollutants measured in areas with intensive RSOB activities suggest a high exposure risk and potential adverse health effects. The fact that agricultural land in Asia is widely distributed in suburban and rural areas, where people live and work, further intensifies the exposure risk. In some areas, such as the Red river delta of Vietnam, multiple small and short-lived RSOB fires can be densely seen during the harvesting months, hence, can seriously deteriorate air quality [36]. The emissions are widely dispersed to cause high air pollution levels not only in the rural areas where RSOB occurs but also in the adjacent urban areas [27, 37]. Furthermore, RSOB activity is more intensively practiced in the dry season when air pollution is already high due to the stagnant meteorological conditions and thus exaggerates haze episodes [4, 37].

Presently, the impacts of CROB including RSOB are often overlooked in many SEA countries and receive less concern

from society as compared to catastrophic transboundary haze caused by forest fires. The negative impacts of RSOB, such as the effects of smoke on human health, should be widely disseminated to raise awareness and thereby encourage farmers to use non-open burning alternatives for crop residue management.

Non-burning alternatives for rice straw management

Non-burning alternatives: rice straw is a valuable resource that should be recovered rather than disposed of by open burning. There are non-open burning alternatives including off-site uses of RS as medium for mushroom cultivation, animal feed and bedding, garden mulching, or composting. Further, RS can be converted into biochar, processed fuel such as bioethanol or briquettes/pellets, and building materials [38, 39]. However, the labor and cost for collection and transportation of bulky loose RS remain a challenge. Other constraints exist, for example, the presence of a high silica (Si) content in RS affects the digestion capability of the livestock. The traditional uses of RS in handicraft making (hats, mats, and decoration items) or as construction material could be promoted but need suitable business models to sustain.

A promising alternative of "ploughing for on-site degradation" has been promoted in Thailand. Accordingly, the harvested paddy is ploughed using a powerful machine to incorporate RS into the soil following with the application of water, bio-extracted liquid, and bio-fertilizer to accelerate the degradation. However, local farmers do not generally prefer the method primarily because it still requires a considerable amount of time to make the paddy soil ready for the next crop plantation. In addition, the high cost of such a ploughing machine is also an issue. Therefore, most farmers still prefer to continue RS field burning activities. However, the results of a survey showed that the local farmers are indeed aware of the negative effects of RS field burning on paddy soil quality, e.g., the soil structure becomes hard after RSOB and some organic nutrient substances on the topsoil are burned to ash. The effects of smoke on human health are not yet recognized while those on the safety of on-road transport due to reduced visibility were of concern [40].

Cooking with loose RS in a simple tripod cookstove is traditionally practiced in rural areas of Asia [41, 42]. However, this low efficient cooking system consumes large amounts of fuel and generates huge amounts of emissions that affect both indoor and outdoor air quality. Some densification techniques can be applied to produce RS derived solid fuels that include, for example, roped/bundled straw, briquettes, and pellets, which have higher fuel density and are easier to store and transport than loose RS. Further, clean cookstoves can be used to effectively burn RS-derived fuel while generating less emissions. However, these conversion technologies are not yet fully developed or adapted for RS. Overall, the thermochemical conversion of RS into bioenergy is still not popularly applied

due to several technical constraints associated with RS's chemical and physical properties like high ash and silica content, rigorous crystalline structure, and low bulk density [42-44]. Combustion is the most mature technology used to generate energy from RS, but it also encounters several major shortcomings including corrosion/fouling of equipment due to high silica and potassium/alkali contents, ash accumulation and slagging, among others [44, 45].

Production of rice straw pellet fuel for cooking: making RS pellets (pelletization) is one densification method to produce RS-derived solid fuel. This process can significantly increase the bulk density of RS, i.e. up to 600-700 kg/m³ from, for example, 60-90 kg/m³ of baled RS [46]. To improve the quality of the produced pellets, RS is usually mixed with woody/bamboo biomass and/or other additives. In particular, woody/bamboo biomass with its high lignin content and low ash content enhances the energy content and improves pellet durability while reducing its inorganic content [47].

The air quality team at the Asian Institute of Technology (AIT) has conducted research on RSOB emissions and assessed the associated impacts on air quality and climate forcing during the last 20 years. The AIRPET (Air pollution research project and network) from 2000-2010, sponsored by Swedish International Development Cooperation Agency (SIDA) and coordinated by AIT in collaboration with 6 Asian national research partners from China, India, Indonesia, Philippines, Thailand, and Vietnam [48], started emission characterization and modeling studies related to RSOB in SEA [49].

Recognizing RS as a valuable resource, the AIT team worked to identify alternatives that recover energy from this agricultural waste allowing farmers to commoditize waste that is prone to burning. In the Partnerships for Enhanced Engagement in Research (PEER) - SEA project "Assessment of impacts of the emission reduction measures of short-lived climate forcers on air quality and climate in SEA" (2012-2016) sponsored by U.S. Agency for International Development (USAID), the team quantified the RSOB emissions in SEA and assessed the impacts on air quality and climate forcing using the modeling tool [50]. A spin-off project under the Sustainable Mekong Research Network (SUMERNET) phase 3, sponsored by SIDA and titled "Turning rice straw into cooking fuel for air quality and climate co-benefit in selected Greater Mekong Subregion countries" (2016-2018), was conducted in cooperation with the Energy program at AIT and research partners in Thailand, Vietnam, and Cambodia to examine several options to turn RS into cooking fuel like RS bundles, briquettes, and pellets.

A laboratory scale pelletizing machine was developed in this project [51] and successfully produced RS pellets that can be burned effectively in a Mimimoto gasifier cookstove (GCS) (<https://www.engineeringforchange.org/solutions/product/mimi-moto/>). The GSC-pellet cooking system was



Fig. 3. (A) A pelletizing machine in an agricultural machinery company in Hanoi, (B) RS pellets produced, and (C) GCS-pellet cooking system demonstrated in Hanoi [52].

demonstrated at project sites in Cambodia, Thailand, and Vietnam and gained a general acceptance from farmers. A few shortcomings have been documented such as the strongly sintered ash remaining in GSC after pellet burning, which was difficult to remove from the stove and the ash material was found to be too hard for use for soil conditioning.

A supplementary award for PEER-SEA was provided to AIT for translating evidence-to-action in a demonstration project "Technology acceleration to transfer rice straw derived fuel and gasifier cookstove in Vietnam" (2017-2018). The project team successfully produced RS pellets using a full-scale prototype pelletizing machine in cooperation with the local project partners at the Hanoi National University and an agricultural machinery company in Hanoi (Fig. 3). The pellets burned well in the selected Mimimoto GCS without visible smoke. Certain modifications of feeding materials and pelletizing technical conditions make the ash soft enough to be removed easily from the stove and also to apply directly on soil. This cooking system has a high thermal efficiency, hence, consumes less fuel for cooking a meal. The emission measurements showed that the amount of PM_{2.5} emitted from the cookstove when burning 1 kg of RS pellets was only about one-fifth (1/5) of that from RSOB pile burning. The RS pellets can be used for domestic cooking and are even more relevant for commercial cooking to substitute, for example, polluting honeycomb coal briquettes. Thus, this would reduce exposure to both indoor and outdoor air pollution and provide great health benefits.

The production of RS pellets and the selection of the optimal cookstove for burning the fuel create an opportunity to meaningfully recover this valuable agricultural waste and, at the same time, create an income source for farmers through selling RS and/or RS pellets. It provides an alternative to reduce RSOB and brings in benefits of clean air and climate though emission reduction. At the same time, RS pellet production helps cut down the consumption of fossil fuel (to reduce climate impacts) and wood fuel (to save trees), hence, providing multiple benefits.

A complete RS grinding-pelletizing machine should be further developed and demonstrated to bring the technology closer to end-users. Modifications to feeding material mixture compositions and the pelletizing technical conditions may be

exploited to produce pellets for other purposes like animal feedstock, organic fertilizers, and soil conditioners. The demand, willingness-to-pay by users, cost-benefit analysis, and potential environmental impacts should be analysed. Business models may be developed that involve participation of the private sector to produce and market RS pellets.

Conclusions

RSOB releases huge amounts of toxic air pollutants that seriously deteriorate local air quality not only in populated rural areas but also in nearby cities. Intensive RSOB in the dry period, when stagnant atmospheric conditions prevail, intensifies the formation of haze episodes. High levels of toxic and carcinogenic compounds in the ambient air during intensive burning periods indicate high exposure risks and potential adverse health effects.

Several non-burning alternatives are available to RSOB but certain constraints exist. The production of RS pellets for cooking fuel is a promising approach to minimize RSOB activities and gain multiple benefits. Further studies should include a detailed cost-benefit analysis of the application and should develop practical guidelines for the production of RS pellets with suitable physical and thermal characteristics of the fuel.

For successful elimination of RSOB in SEA, beside the technology development, formulation and implementation of appropriate policies should be in place to mobilise participation from all stakeholders. A strict “ban” on RSOB alone may not work effectively, but the enforcement should be done along with providing suitable and workable alternatives with subsidies/incentives. Negative impacts of RSOB, specifically the effects of smoke on human health, should be widely disseminated to raise awareness and thereby encourage farmers to opt for non-open burning alternatives for rice straw management. The benefits brought about by non-open burning scenarios to ambient air quality and health should be demonstrated in future studies by using a detailed emission inventory and dispersion modelling approach. Suitable business models involving the private sector should be developed that incorporate sufficient incentives to encourage farmers to stop open burning.

ACKNOWLEDGEMENTS

The author, also the PI of the regional projects, would like to specially thank the sponsors of Swedish International Development Cooperation Agency for the generous funding supports to the AIRPET and SUMERNET projects, and U.S. Agency for International Development for the support to the SEA-PEER project with supplemental awards. The cooperation of AIT colleagues and national research partners from Thailand, Vietnam, Indonesia, and Cambodia, as well as the local company in Hanoi is highly acknowledged. The project activities would not be possible without the active participation from AIT students and project research staff.

COMPETING INTERESTS

The author declares that there is no conflict of interest regarding the publication of this article.

REFERENCES

- [1] N.T. Kim Oanh, D.A. Permadi, P.A. Salam, P.K. Lieu, D.V. Hieu, P. Pongkiatkul, K. Kositkanawuth, K. Sothea, C. Elyan, P. Hopke, C.T. Hoanh (2019), “Assessment of co-benefits of using rice straw derived solid fuel for cooking to reduce emissions of agro-residue open burning in selected GMS countries”, *Chapter 8, Development and Climate Change in the Mekong Region: Case Studies*, SIRD Publishing Co., Malaysia, 349pp.
- [2] K. Anezaki, N. Kashiwagi (2021), “Daily variations and factors of atmospheric PCDD/Fs in post-harvest paddy fields: PCDD/F source estimation using a Bayesian semi-factor model”, *Chemosphere*, **268**, DOI: 10.1016/j.chemosphere.2020.129292.
- [3] Chau-Thuy Pham, Yaowatat Boongla, Trung-Dung Nghiem, Huu-Tuyen Le, Ning Tang, Akira Toriba, Kazuichi Hayakawa (2019), “Emission characteristics of polycyclic aromatic hydrocarbons and nitro-polycyclic aromatic hydrocarbons from open burning of rice straw in the north of Vietnam”, *Int. J. Environ. Res. Public Health*, **16**(13), pp.2343-2360.
- [4] A. Tipayarom, N.T. Kim Oanh (2020), “Influence of rice straw open burning on levels and profiles of semi-volatile organic compounds in ambient air”, *Chemosphere*, **243**, DOI: 10.1016/j.chemosphere.2019.125379
- [5] N.T. Kim Oanh, D.A. Permadi, P.K. Hopke, R.K. Smith, N.P. Dong, D.A. Nguyen (2018), “Annual emissions of air toxics emitted from crop residue open burning in Southeast Asia over the period of 2010-2015”, *Atmospheric Environment*, **187**, pp.163-173.
- [6] N.T. Kim Oanh, A. Tipayarom, L.B. Thuy, D. Tipayarom, C.D. Simpson, D. Hardie, L.J. Sally Liu (2015), “Characterization of gaseous and semi-volatile organic compounds emitted from field burning of rice straw”, *Atmospheric Environment*, **119**, pp.182-191.
- [7] X. Zhang, Y. Lu, Q.G. Wang, X. Qian (2018), “A high-resolution inventory of air pollutant emissions from crop residue burning in China”, *Atmospheric Chemistry and Physics*, DOI: 10.5194/acp-2017-111.
- [8] J. Jimenez, C. Claiborn, R. Dhammapala, C.D. Simpson (2007), “Developing a source fingerprint for burning of wheat and Kentucky bluegrass stubble in Eastern Washington and Northern Idaho”, *Environmental Science and Technology*, **41**(22), pp.7824-7829.
- [9] A.H. Miguel, A.E.F. Fernandez, P. Jaques, J.R. Froines, B.L. Grant, P.R. Mayo, C. Sioutas (2004), “Seasonal variation of the particle size distribution of polycyclic aromatic hydrocarbons and of major aerosol species in Claremont, California”, *Atmospheric Environment*, **38**(20), pp.3241-3251.
- [10] M.D. Hays, P.M. Fine, C.D. Geron, M.J. Kleeman, B.K. Gullett (2005), “Open burning of agricultural biomass: physical and chemical properties of particle-phase emissions”, *Atmospheric Environment*, **39**(36), pp.6747-6764.
- [11] B. Gullett, A. Touati (2003), “PCDD/F emissions from burning wheat and rice field residue”, *Atmospheric Environment*, **37**(35), pp.4893-4899.
- [12] M.O. Andreae, P. Merlet (2001), “Emission of trace gases and aerosols from biomass burning”, *Global Biogeochemical Cycles*, **15**(4), pp.955-966.
- [13] B.M. Jenkins, A.D. Jones, S.Q. Turn, R.B. Williams (1996), “Particle concentrations, gas-particle partitioning, and species intercorrelations for polycyclic aromatic hydrocarbons (PAH) emitted during biomass burning”, *Atmospheric Environment*, **30**(22), pp.3825-3835.
- [14] I.A.S. Hussein, M.S.M. Mansour (2016), “A review on polycyclic aromatic hydrocarbons: source, environmental impact, effect on human health and remediation”, *Egyptian Journal of Petroleum*, **25**(1), pp.107-123.
- [15] E.J. Mrema, F.M. Rubino, G. Brambilla, A.M. Aristidis, M. Tsatsakis, C. Colosio (2013), “Persistent organochlorinated pesticides and mechanisms of their toxicity”, *Toxicology*, **307**, pp.74-88.
- [16] IPCC (2013), *Climate Change 2013: the Physical Science Basis*, Cambridge University Press, United Kingdom and New York, USA, 1535pp.

- [17] Ngo Thanh Danh, Lai Nguyen Huy, Nguyen Thi Kim Oanh (2016), "Assessment of rice yield loss due to exposure to ozone pollution in Southern Vietnam", *Science of the Total Environment*, **566-567**, pp.1069-1079.
- [18] N.T. Kim Oanh, L.B. Thuy, D. Tipayarom, B.R. Manadhar, P. Pongkiatkul, C.D. Simpson, L.J. Sally Liu (2011), "Source characterization of aerosol emission from field burning of rice straw", *Atmospheric Environment*, **45(2)**, pp.493-502.
- [19] K. Lasko, K. Vadrevu (2018), "Improved rice residue burning emissions estimates: accounting for practice-specific emission factors in air pollution assessments of Vietnam", *Environmental Pollution*, **236**, pp.795-806.
- [20] T.J. Christian, B. Kleiss, R.J. Yokelson, R. Holzinger, P.J. Crutzen, W.M. Hao, B.H. Saharjo, D.E. Ward (2003), "Comprehensive laboratory measurements of biomass-burning emissions: 1. Emissions from Indonesian, African, and other fuels", *Journal of Geophysical Research*, **108(D23)**, pp.4719-4732.
- [21] H. Zhang, X. Ye, J. Cheng, X. Yang, L. Wang, R. Zhang (2008), "A laboratory study of agricultural crop residue combustion in China: emission factors and emission inventory", *Atmospheric Environment*, **42(36)**, pp.8432-8441.
- [22] H. Keshtkar, L.L. Ashbaugh (2007), "Size distribution of PAH particulate emission factors from agricultural burning", *Atmospheric Environment*, **41(13)**, pp.2729-2739.
- [23] E. Sanchis, M. Ferrer, S. Calvet, C. Coscoll, V. Yus, M. Cambra-Lopez (2014), "Gaseous and particulate emission profiles during controlled rice straw burning", *Atmospheric Environment*, **98**, pp.25-31.
- [24] G. Cao, X. Zhang, S. Gong, F. Zheng (2008), "Investigation on emission factors of particulate matter and gaseous pollutants from crop residue burning", *Journal of Environmental Sciences*, **20(1)**, pp.50-55.
- [25] R.J. Sheesley, J.J. Schauer, Z. Chowdhury, G.R. Cass, B.R.T. Simoneit (2003), "Characterization of organic aerosols emitted from the combustion of biomass indigenous to South Asia", *Journal of Geophysical Research*, **108(D9)**, pp.4285-4300.
- [26] T. Korenaga, X. Liu, Z. Huang (2001), "The influence of moisture content on polycyclic aromatic hydrocarbons emission during rice straw burning", *Chemosphere Global Change Science*, **3(1)**, pp.117-122.
- [27] D. Narita, N.T. Kim Oanh, K. Sato, M. Huo, D.A. Permadi, N.N. Ha Chi, R. Tanatat, I. Pawarmart (2019), "Pollution characteristics and policy actions on fine particulate matter in a growing Asian economy: the case of Bangkok metropolitan region", *Atmosphere*, **10(5)**, pp. 227-245.
- [28] N.T. Kim Oanh (2013), "Integrated approach to rice straw management for reduction of field burning activity", *Integrated Air Quality Management: Asian Case Studies*, CRC Press Boca Raton, USA, 424pp.
- [29] K. Ravindra, T. Singh, S. Mor (2019), "Emissions of air pollutants from primary crop residue burning in India and their mitigation strategies for cleaner emissions", *Journal of Cleaner Production*, **208**, pp.261-273.
- [30] M. Viana, J.M. López, X. Querol, A. Alastuey, D. García-Gacio, G. Blanco-Heras, P. López-Mahía, M. Piñeiro-Iglesias, M.J. Sanz, F. Sanz, X. Chi, W. Maenhaut (2008), "Tracers and impact of open burning of rice straw residues on PM in Eastern Spain", *Atmospheric Environment*, **42(8)**, pp.1941-1957.
- [31] C.F. Wu, J. Jimenez, C. Claiborn, T. Gould, C.D. Simpson, T. Larson, L.J. Sally Liu (2006), "Agricultural burning smoke in eastern Washington: part II: exposure assessment", *Atmospheric Environment*, **40(28)**, pp.5379-5392.
- [32] H.H. Yang, C.H. Tsai, M.R. Chao, Y.L. Su, S.M. Chien (2006), "Source identification and size distribution of atmospheric polycyclic aromatic hydrocarbons during rice straw burning period", *Atmospheric Environment*, **40(7)**, pp.1266-1274.
- [33] K. Torigoe, S. Hasegawa, O. Numata, S. Yazaki, M. Matsunaga, N. Boku, H. Makoto, H. Ino (2000), "Influence of emission from rice straw burning on bronchial asthma in children", *Pediatrics International*, **42(2)**, pp.143-150.
- [34] UNEP-WMO (2011), *Integrated Assessment of Black Carbon and Tropospheric Ozone*, World Meteorological Organization and United Nations Environment Programme, 285pp.
- [35] D. Shindell, et al. (2012), "Simultaneously mitigating near-term climate change and improving human health and food security", *Science*, **335(6065)**, pp.183-189.
- [36] K. Lasko, K.P. Vadrevu (2018), "Biomass burning emissions variation from satellite derived land cover, burned area, and emission factors in Vietnam", *Land Atmospheric Research Applications in South and Southeast Asia*, pp.171-201, DOI: 10.1007/978-3-319-67474-2_9.
- [37] D. Tipayarom, N.T. Kim Oanh (2007), "Effects from open rice straw burning emission on air quality in the Bangkok metropolitan region", *ScienceAsia*, **33(3)**, pp.339-345.
- [38] E.B. Belal (2013), "Bioethanol production from rice straw residues", *Brazilian Journal of Microbiology*, **44(1)**, pp.225-234.
- [39] J.Y. Park, R. Shiroma, M.I. Al-Haq, Y. Zhang, M. Ike, Y. Arai-Sanoh, A. Ida, M. Kondo, K. Tokuyasu (2010), "A novel lime pretreatment for subsequent bioethanol production from rice straw-calcium capturing by carbonation (CaCCO) process", *Bioresource Technology*, **101(17)**, pp.6805-6811.
- [40] K. Kanokkanjana, A. Bridhikitti (2007), "Sustainable rice straw management for urban air pollution reduction in Bangbuathong, Nonthaburi province, Thailand", *Final Report of the CIDA-AIT Partnership*, Alumni demonstration project, Pathumthani Asian Institute of Technology, 23pp.
- [41] Lai Nguyen Huy, Nguyen Thi Kim Oanh, Nguyen Hong Phuc, Chu Phuong Nhung (2020), "Survey-based inventory for atmospheric emissions from residential combustion in Vietnam", *Environmental Science and Pollution Research*, **28**, pp.10678-10695.
- [42] N.N.H. Mai (2015), *Assessment of the Impact of Using Rice Straw Derived Cooking Fuel on the Emission of Air Pollution and Climate Forcers*, Master's Thesis, Asian Institute of Technology, 101pp.
- [43] Y. Zhang, et al. (2012), "Physical properties of rice residues as affected by variety and climatic and cultivation conditions in three continents", *American Journal of Applied Sciences*, **9(11)**, pp.1757-1768.
- [44] M. Rozainee, S. Ngo, A. Salema, K. Tan, M. Ariffin, Z. Zainura (2008), "Effect of fluidising velocity on the combustion of rice husk in a bench-scale fluidised bed combustor for the production of amorphous rice husk ash", *Bioresource Technology*, **99(4)**, pp.703-713.
- [45] M.C. Maguyon-Detras, M.V.P. Migo, N. Van Hung, M. Gummert (2020), "Thermochemical conversion of rice straw", *Sustainable Rice Straw Management*, pp.43-64, DOI: 10.1007/978-3-030-32373-8_4.
- [46] N.V. Hung, M.C. Maguyon-Detras, M.V. Migo, R. Quilloy, C. Balingbing, P. Chivenge, M. Gummert (2020), "Rice straw overview: availability, properties, and management practices", *Sustainable Rice Straw Management*, pp.1-13, DOI: 10.1007/978-3-030-32373-8_1.
- [47] Z. Liu, L. Xing'e, F. Benhua, J. Zehui, C. Zhiyong, Y. Yan (2013), "The properties of pellets from mixing bamboo and rice straw", *Renewable Energy*, **55**, pp.1-5.
- [48] N.T. Kim Oanh, N. Upadhyay, Y.-H. Zhuang, Z.-P. Hao, D.V.S. Murthy, P. Lestari, J.T. Villarin, K. Chengchua, H.X. Co, N.T. Dung, E.S. Lindgren (2006), "Particulate air pollution in six Asian cities: spatial and temporal distributions, and associated sources", *Atmospheric Environment*, **40(18)**, pp.3367-3380.
- [49] N.T. Kim Oanh (2014), *Improving Air Quality in Asian Developing Countries: Compilation of Research Findings*, NARENCA, 435pp.
- [50] D.A. Permadi, N.T. Kim Oanh, R. Vautard (2018), "Assessment of emission scenarios for 2030 and impacts of black carbon emission reduction measures on air quality and radiative forcing in Southeast Asia", *Atmospheric Chemistry and Physics*, **18**, pp.3321-3334.
- [51] D.N. Cuong (2016), *A Study on Production and Combustion Behavior of Rice Straw Pellets*, Master's Thesis, Asian Institute of Technology, 140pp.
- [52] PEER (2018), "Evidence-to-action supplements (sub-award)", *Assessment of Impacts of the Emission Reduction Measures of Short-Lived Climate Forcing Pollutants on Air Quality and Climate in Southeast Asia*, USAID/NAS, 30pp.