

Composition and floral diversity in Andean grasslands in natural post-harvest restoration with *Lepidium meyenii* Walpers

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ABSTRACT

The Andean grassland ecosystems undergo natural and anthropogenic degradation processes. The change of land use for agricultural use is the greatest threat, with a great loss of biodiversity followed by a very slow process of revegetalization. The objective was to assess the richness, abundance and diversity, alpha and beta, in areas of two-, three-, five-, six- and eight years of post-harvest abandonment of Lepidium meyenni Walpers. Ten affected areas were selected for agrostological evaluation, through four linear transects of interception points with 100 records each, applied on the edge and inside the affected area, inside and outside the affected areas, as well as samples composed of soils for each area. Composite soil samples were collected from each transect and study area to analyze their physical and chemical properties. The data were analyzed using the generalized linear mixed model with Rstudio v 5.3.2, and the multivariate analysis of canonical correspondence between biological and environmental variables, using CANOCO v.1.4 software. A significant difference was found between floral composition and alpha diversity according to abandonment time and seasonal periods. The greatest richness and abundance was observed between five- and eight years of abandonment, due to the greater presence of perennial species. Linear correspondence of alpha richness and diversity with organic matter, nitrogen, soil phosphorus and abandonment time was observed.

Keywords: alpha diversity, Andean grassland, beta diversity, dominance richness.

Composição e diversidade florística em campos andinos em restauração pós-colheita natural com *Lepidium meyenii* Walpers

RESUMO

Os ecossistemas de prados andinos sofrem processos de degradação natural e antropogênica. A mudança de uso da terra para uso agrícola é a maior ameaça com grande perda de biodiversidade, seguido por um processo muito lento de revegetação. O objetivo do estudo foi avaliar a riqueza, abundância e diversidade alfa e beta em áreas de dois, três, cinco, seis e oito anos pós-colheita *Lepidium meyenni* Walpers. Foram estabelecidas dez áreas afetadas para avaliação agrostológica por meio de quatro transectos lineares de pontos de interceptação com 100 registros aplicados na borda e no interior da área afetada. Amostras compostas de solo foram coletadas de cada transecto e da área de estudo para analisar suas propriedades físicas e químicas. Os dados foram analisados por meio do modelo linear generalizado misturado com



This is an Open Access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original work is properly cited. Rstudio v 5.3.2 e da análise multivariada de correspondência canônica entre variáveis biológicas e ambientais, utilizando o software CANOCO v.1.4. Uma diferença significativa foi encontrada entre composição florística e diversidade alfa de acordo com o tempo de abandono e períodos sazonais. A maior riqueza e abundância foi observada entre cinco e oito anos de abandono, devido à maior presença de espécies perenes. Observou-se correspondência linear de riqueza e diversidade alfa com matéria orgânica, azoto, fósforo do solo e tempo de abandono.

Palavras-chave: diversidade alfa, diversidade beta, pastagem Andina, riqueza florística e dominância.

1. INTRODUCTION

Natural grassland ecosystems include a diversity of plant species and provide habitat for diverse wildlife species, as well as important resources for humanity and the environment (Rolando *et al.*, 2017). The floral diversity of the grassland ecosystems of these high Andean ecoregions (Britto, 2017) is similar in complexity and heterogeneity to the patterns of floral variation determined by gradients (Cuesta and Becerra, 2012). Species richness generally declines as elevation increases (McCain and Grytnes, 2010). Whittaker (1960) demonstrated a positive relationship between temperature and precipitation and species richness for many diversity patterns (Hawkins *et al.*, 2003) and as the study area increases, the number of species found increases (Rosenzweig, 1995). On the other hand, the history of agricultural-livestock use (Mudongo *et al.*, 2016) and the feeding ofanimal by grazing (Yaranga *et al.*, 2018), the strategy of grasslands' adaptation to climate change and the economic factors of the herders (Hinojosa *et al.*, 2016) determine the diversity of species.

These ecosystems are being severely disturbed in their structure and function by the change of pasture land use for agricultural purposes, with a great loss of biodiversity and the generation of a secondary succession process (Schmidt, 2018). Revegetation begins with the establishment of pioneering invasive species unrelated to the original composition (Wiesmair *et al.*, 2017), generating important changes in physiognomic and physiological properties (Gartzia *et al.*, 2016; Silva Mota *et al.*, 2018) and causing the decrease of multiple environmental services according to the variation of spatiotemporal scenarios (Tang *et al.*, 2018). In different parts of the world faced with these problems, strategies have been developed to reverse these processes through conservation and restoration (Jurado *et al.*, 2013; Caro *et al.*, 2014).

In the central Andes of Peru, the change in the use of Andean grasslands through the cultivation of *Lepidium meyenii* Walpers, due to the removal of the soil with loss of vegetation cover, constitutes the greatest regional concern. This is due to the fact that the recovery of plant communities is slow during the four years of rest, and their capacity to shelter diverse species is diminished (Cao *et al.*, 2017), and they are gradually being repopulated by herbaceous communities very different from the original ecosystem, substituting the original plant community for another (Cárdenas, 2013).

The quantification of species richness, abundance and similarity through alpha and beta diversity indices allows studying changes in the plant community. For this, it is important to identify the species with the highest importance (Silva Mota *et al.*, 2018) based on the relative abundance and community structure of species (Moreno, 2001; Moreno *et al.*, 2011; Barragán Solís *et al.*, 2011), since these determine the characteristics of the plant community and the function of the ecosystem according to environmental gradient and limiting factors referred to by Whittaker and Clements (Hagen, 2012), in addition to being an indication of the level of relevance among the other species and the level of spatial occupation (Enriquez *et al.*, 2016). Alpha diversity, also known as specific diversity, varies in time from one site to another, in relation to the spatial and temporal gradient between communities or habitats (Halffter *et al.*, 2005), and is used in the ecological evaluation of conservation and restoration of ecosystems



(Caballero-Vázquez and Vega-Cendejas, 2012). Beta diversity, on the other hand, implies knowing the degree of replacement of species between communities or the number of species that share two or more communities (Moreno, 2001).

This study assessed the richness, abundance and floral diversity of vascular plants in Andean grassland areas with land-use change, in space-time gradients of two-, three-, five-, six- and eight years of post-harvest abandonment of *Lepidium meyenii* Walpers and two seasonal periods.

2. MATERIAL AND METHODS

2.1. Study area

The study was conducted in "Lomo Largo", in the district of Pomacancha, province of Jauja and department of Junín, between coordinates 11°33'37.50" and 11°32'02.35" and between altitudes 3998 and 4262 meters (Figure 1). The study area is covered with natural grass vegetation, mostly poaceae and asteraceae, with a scarce presence of shrubs such as *Margiricarpus pinnatus* and *Chuquiraga spinosa*, which are generally observed in rocky areas and rocky outcrops. The soils are of colluvial origin with parental material formed mainly by rocks: sandstones, shales, slates and quartzite; of very variable depth between 10 and 30 cm, and generally poor in nutrients. The soil is classified as "entisol" according to the "Soil Taxonomy" of the USA, containing less than 30% of rock fragments (USDA, 2018). However, due to the high physiographic heterogeneity, other soils such as inceptisols and mollisols are observed, due to the altitude of over 3800 meters, cold temperatures, grass cover, and slopes differentiated between 2% and 15%.



Figure 1. Location map of the study. Área 1 starts on the north side.

The rainy season occurs from October to April, with a peak of greater precipitation between January and March, with accumulated precipitation between 106 and 118 mm per month. The dry season occurs from May to September with minimum precipitation of 8.9 to 28



mm per month. The average maximum daytime temperature fluctuates between 12.87 and 14.63°C, the minimum nighttime and early morning temperature fluctuates between 0.8 and 5.4°C, according to data obtained from the Ricran Meteorological Station.

2.2. Methods

2.2.1. Agrostological evaluation and soils

Ten areas affected by land-use change were identified and established with two-, three-, five-, six- and eight years of post-harvest abandonment with *L. meyenii*, based on cultivation history provided by the ranchers at the study site. The agrostological evaluation was carried out using the "point interception line" method proposed by the Gloria Method (Enriquez *et al.*, 2016), which consisted of reading 100 points at intervals of 10 cm, in which the following were recorded: vascular plant species, mulch, moss, eroded soil and rock. In each affected area, four transects on the outer edge (native vegetation) and four transects inside the affected area (revegetation) were carried out in May (rainy season) and August (dry season) in 2017, according to the protocol of the GLORIA methodology (Cuesta and Becerra, 2012), used by the Network to monitor the impacts of climate change on biodiversity in the countries of the Andean region (Halloy, 2011). Species recognition was carried out by experts from the Faculty of Zootechnics and the Faculty of Forestry and Environmental Sciences (depository of the constructed herbarium) of the National University of Central Peru, authorized by Resolution No. 095-2018-MINAGRI-SERFOR/DGGSPFF of the Forestry and Wildlife Service.

Soil sampling was carried out according to the technique recommended by the Ministry of the Environment [MINAM] (Perú, 2014) at a depth of 30 cm. Three soil samples were collected per transect, which were mixed to obtain a sample composed of native vegetation and another from the revegetation area, which were sent to the soil laboratory of the National Agricultural Research Institute (INIA) Santa Ana de Huancayo, for the analysis of soil texture, pH, organic matter (OM), nitrogen (N), phosphorus (P) and potassium (K).

2.2.2. Data analysis

The data obtained in the field and laboratory were systematized in a double entry matrix in Excel sheet, adopting the method of the species list presentation used by Uchida et al. (2018) for each area, separated between native vegetation and revegetalization area. The agrostological sampling quality test was performed using the cumulative curve method recommended by Kindt and Coe (2011). The floral data was transformed into abundance using the formula: number of individuals of the species divided by the total number of individuals recorded for each transect. The procedure performed by Ali et al. (2017) in the Significant Minimum Difference (LSD) comparison test was adopted to separate the different means ($p \le 0.05$) using the generalized mixed linear model, with the use of RStudio software version 3.5.2. in the analysis of richness and abundance, Shannon Wiener alpha diversity index (H'), and Jaccard (Ij) beta diversity index (Moreno, 2001). Multivariate analysis was also carried out between biological variables and environmental variables (revegetation time, seasonal period and soil characteristics) as explanatory variables, by means of the canonical correspondence analysis (CCA) that allows for visualization in a geometric figure, the proximities between the set of biological data with certain predictor variables (Cayuela, 2011) in a final biplot graph (Montanero, 2018), using CANOCO software version 4.5.

3. RESULTS AND DISCUSSION

3.1. Richness and abundance

In the native vegetation around the affected areas, 24 families with 57 genera and 79



species were found, of which the family Poaceae contributes with 28 species, the family Asteraceae with 16 species, followed by Fabaceae with four species. In the areas in post-harvest revegetation of *L. meyenii*, abandoned after the only harvest of the product, 23 families with 55 genera and 73 species were found, including 23 species of the family Poaceae and 16 species of the family Asteraceae. The composition of genera and species in both vegetations varied according to seasonal periods and post-harvest abandonment time. The greatest importance of the Poaceae family is a fundamental characteristic of natural grass ecosystems of any environmental condition in the world (Fiallos *et al.*, 2015); it diminishes or loses its composition due to anthropogenic intervention, and does not show a complete recovery, despite several years of revegetation (Sedlar *et al.*, 2017). Significant differences were observed in the number and abundance of species between seasonal periods, mainly in revegetation areas. Richness varied according to seasonal periods and post-harvest abandonment time with *L. meyenii*.

The small variation in the number of families could be due to the environmental conditions of location within the humid Subalpine Tropical paramo life zone, where the 10 evaluated areas are located. However, the variation in the number of genera and species was significant after five years of post-harvest abandonment, due to the rugged microclimatic condition of the soil and the scarcity of rainfall (Uchida *et al.*, 2018). The composition and floristic structure continued until the age of 15, since the process of revegetation in the areas affected by land-use change follows a sequential pattern, first entering temporary species with great capacity to adapt to hostile environmental situations, by the existence of free spaces that allow them to take direct advantage of solar radiation, mainly the infrared that enhances their photosynthetic capacity (Silva Mota *et al.*, 2018), then they are gradually replaced by perennial species more tolerant to competition for light and space. This information is valid for decision making in research processes and the planning of directed restoration actions in grassland ecosystems.

The variation of the richness in the areas of revegetalization according to seasonal period was significant, with p = 0.001 and at post-harvest abandonment time level with *L. meyenii*. The areas of five-, six- and eight years were superior to the first years with p = 0.01, according to the ANOVA analysis using the generalized mixed linear model (Figure 2), the most common species in years two and three post-harvest with *L. meyenii* were: *Paspalum pygmaeum* Hack, *Alchemilla pinnata* Ruiz & Pav, *Bidens andícola* Kunt *and Oxalis sp*, while from the age of five they were: *P. pilgerianum* Hack, *A. pinnata*, *Calamagrostis heterophylla* Wedd, *Stipa depauperata*, *B. andícola*.



Figure 2. variation of the richness in plots in revegetalization, according to (**A**) seasonal period and (**B**) postharvest time of *L. meyeni*.

The abundance in revegetation plots also varied significantly between seasonal periods with p = 0.0001, while at the post-harvest time level no significant difference was found (Figure 3). The 10 species that showed the greatest abundance were: *P. pilgerianum, P. pygmaeum,*



Trifolium amabili Lojac, *A. pinnata, Oxalis andicolae* Knut, *sp, Calamagrostis heterophylla, Paronychia sp, Rumex asetocella, S. depauperata* Pilg and *B. andícola*; of which Alchemilla, Calamagrostis and Stipa are perennial plants, which make a notable presence in the fifth year so that the peak of the curve is also observed.



Figure 3. Upper limit of means of abundance, in areas in revegetation according to (A) seasonal period and (B) time of post-harvest abandonment of *L. meyenii*.

The richness and scarce abundance in the first post-harvest years with *L. meyenii* leaves soils with very little cover, affecting soil conditions mainly by loss of organic matter and nutrients (Mirazadi *et al.*, 2017). The perennial species found in this study are similar to those found in Puno by Cutimbo *et al.* (2016), and the decrease in species coincides with the results in fragmented areas of Condori (2012), which reaffirms that the floral composition in high mountain ecosystems is somewhat similar.

3.2. Floristic diversity

3.2.1. Alpha Diversity Index

The relationship between the number of species and their abundance expressed by H' in native vegetation (Table 1) ranged from 2.6 to 3.0 in the seasonal rainy season and from 1.95 to 2.4 in the dry season. In the revegetation areas, it ranged from 2.37 to 2.89 in the rainy season and 1.35 to 2.46 in the dry season. According to the classification range of the index, these values correspond to mean diversity in native vegetation and low diversity in areas with revegetation (Campo and Duval, 2014; Uchida *et al.*, 2018).

Transects	2 years	3 years	5 years	6 years	8 years
VN - rainy season	2.655	2.638	2.771	2.787	3.057
VN - dry season	2.212	2.015	1.947	2.356	2.111
PR - rainy season	2.374	2.538	2.893	2.811	2.886
PR - dry season	1.713	1.353	2.37	2.449	2.461

Table 1. Alpha diversity indices according to years of abandonment in two seasonal periods.

Legend: 2 years \dots 8 years = years of post-harvest abandonment; VN= native vegetation, PR= areas in revegetalization.

A significant variation of the H' index was observed between seasonal periods with higher value in the rainy period with p = 0.00003. Also a significant increase of the index was observed according to the increase of the post-harvest abandonment time, with more than five-, six- and eight years with p = 0.001 (Figure 4), as opposed to the times of two- and three years. This



indicates that pasture land use change has affected floral diversity for at least five years, and by the relationship between biodiversity and environmental services, particularly the production of useful fodder for livestock (Wehn *et al.*, 2018) which is the economic base of the rural population (Plieninger and Huntsinger, 2018). This behavior of the H' index over time makes it clear that it is important to know the change in species composition over the time of revegetalization in order to understand the effect of anthropogenic action on land-use change in grassland ecosystems (Uchida *et al.*, 2018).



Figure 4. Upper limit of means of diversity of Shannon Wiener and in plots in revegetation according to seasonal period and time of post-harvest abandonment of *L. meyenii*

Several results have been obtained on alpha diversity indices, for the case of H' they vary between 2.4 and 3.36 in Mexico (Álvarez-Lopeztello *et al.*, 2016), 2.56 in Colombia (González-Pinto, 2017), 2.7 to 3.47 in the páramos of Ecuador (Fiallos *et al.*, 2015), in the central Andes of Peru between 2.75 and 3.41 (Yaranga *et al.*, 2018). These values coincide in the classification of medium to low diversity in high mountain pasture conditions, despite its different geographical location on the planet

3.2.2. Beta diversity index

The similarity in the presence of species between pairs of areas in times of abandonment was measured by the Jaccard Index (Table 2), the areas with different post-harvest abandonment times of two-, three-, five-, six- and eight years, compared by pairs, in native vegetation share a greater number of species (12 to 18) in the rainy season than in the dry season (9 to 18); in both cases the degree of similarity is intermediate, with values from 0.31 to 0.5. The areas in revegetation share between 7 to 14 species in the rainy season and 3 to 8 species in the dry season; the degree of similarity varied from intermediate to low. The indices obtained in two-, three- and five- years of post-harvest abandonment showed intermediate levels around 0.5, while in the sixth and eighth years the degree of similarity was low, with values ranging from 0.31 to 0.37 in the rainy season and from 0.29 to 0.46 in the dry season. These results indicated that native vegetation shares between 50% and 38.2% of common species in the seasonal rainy period and between 43.8% and 31% in low water; while in areas with revegetation they share between 61.3% and 38.7% of species in the rainy period between 61.5% and 29.7% in the low-water period.

Beta diversity is a very important ecological indicator that allows us to know the dynamics of the pattern of shared species between areas in recovery according to the years of abandonment, and this is very variable according to the environmental conditions of each place (Condori, 2012; Zhou *et al.*, 2014; Mota *et al.*, 2018). Differences were observed between sampling sites in native vegetation. It is useful to interpret the dynamics of dissimilarity in areas affected by land-use change. The Jaccard index, which measures the capacity to share common species by areas in revegetation, showed higher values than native vegetation, especially in the



first years of revegetation, which was due to the entry of exotic or invasive species adapted to disturbance conditions in the soil (Sedlar et al., 2017), which decreased as the years of postharvest recovery increased, due to the entry of perennial species that replaced the pioneer species. This influenced the observed medium- to low similarity (Uchida et al., 2018). The value of J = 0.79 shown by González-Pinto (2017) is high, compared to the lower indices found in the study, due to the fact that in arid regions the richness of grassland communities is lower than in humid regions (Enriquez et al., 2016; Uchida et al., 2018).

Transects	Years of abandonment	2 years	3 years	5 years	6 years
Native vegitation - rainy season	3	0.500 (16)			
	5	0.532(14)	0.511(13)		
	6	0.444(18)	0.577(14)	0.382(16)	
	8	0.418(12)	0.396(12)	0.464(11)	0.429(13)
Native vegitation – dry season	3	0.438(9)			
	5	0.385(9)	0.375(9)		
	6	0.379(18)	0.486(9)	0.310(10)	
	8	0.375(11)	0.158(11)	0.375(10)	0.371(10)
Revegetation – rainy season	3	0.529(8)			
	5	0.544(7)	0.531(8)		
	6	0.448(10)	0.446(9)	0.408(14)	
	8	0.613(7)	0.420(9)	0.387(13)	0.474(12)
Revegetation – dry season	3	0.615(3)			
	5	0.471(5)	0.500(3)		
	6	0.421(6)	0.500(4)	0.400(7)	
	8	0.600(5)	0.379(5)	0.297(8)	0.463(8)

Table 2. Jaccard similarity index of native vegetation and plots with revegetation, number of common species in parentheses.

3.3. Canonical correspondence (CCA) between species dominance and floristic diversity with soil characteristics, according to post-harvest abandonment time, in plots in natural revegetalization

The analysis of the canonical correspondence between the dominance of species in areas in natural revegetation with the physical-chemical characteristics of the soil, the post-harvest abandonment time of L. meyenii, showed the sensitivity among the species of greatest presence in the five-, six- and eight years of post-harvest abandonment of L. meyenii, with organic matter, nitrogen and phosphorus, both in the seasonal rainy period and in the dry season (Figure 5A). The interest of knowing, in simplified way, the relation of the complex set of environmental data that explain the variation of biological indicators (Cayuela, 2011) has taken into account the diversity of grassland ecosystem species as a key component (Fattahi et al., 2017), which varies according to the sensitivity of the species to the variables of the temporal gradient (years of abandonment) and soil factors, which was the interest of our analysis.

This correspondence confirmed the close relationship between soil and vegetation (Fattahi et al., 2017), which, as a result of the increasingly prolonged abandonment time, shows the correspondence of perennial species with organic matter and nitrogen, which are the most important elements in the quality of soils that energize microbial activity (Ali et al., 2017), and

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plant nutrition (Wang *et al.*, 2017), favored by the increase in biomass produced, litterfall, root death and the increase in microorganisms, arthropods and others, which improve soil quality (Acevedo and Delibes-Mateos, 2013; Niemandt and Greve, 2016). This indicated that the post-harvest vegetative recovery time is a determining factor in the variation and correspondence of soil characteristics with the richness and abundance of species. This correspondence behavior explained the variability of invasive and temporary species in recently abandoned areas that are not demanding in soil quality (Andrade *et al.*, 2015; Sedlar *et al.*, 2017) and the stability of perennial species in the medium- and long term (Gartzia *et al.*, 2016; Mota *et al.*, 2018).



(B) Canonical correspondence of the alpha diversity with the physical-chemical characteristics of the soil and the time of abandonment

Figure 5. Canonical correspondence between A) richness and abundance, B) alpha diversity, in revegetation plots, according to seasonal period and post-harvest abandonment time of *L. meyenii*.

In the analysis of the canonical correspondence of alpha diversity (Shannon Wiener index - H') with the physical-chemical characteristics of the soil, the post-harvest abandonment time of L. meyenii showed a weak linear relationship of the H' index with the organic matter and the abandonment time in the seasonal rainy period, which was not maintained in the low-water period; on the other hand, the affected areas in different abandonment times were concentrated very close to the axes centroid, which indicates that they maintain some independence of the correspondence with other variables. This is due to the fact that the index comes from the application of algorithms to the number of species found (Halffter *et al.*, 2005), so searching for correspondence of the diversity index will not be of practical use.

4. CONCLUSION

It is the Poaceae and Asteraceae families that characterized the high Andean grassland ecosystem that maintained dominance in revegetalization areas, despite the secondary



succession process with invasive species. The post-harvest abandonment time of L. meyenii and the seasonal period affected the behavior of composition, richness and dominance, which determined the variation of alpha and beta diversity. The visible recovery time of areas under revegetalization was evidenced after five years of abandonment, which indicates that the effect of land-use change was very severe in the loss of floral diversity; however, the linear canonical correspondence between floral composition and physical-chemical factors of the soil indicate the possibility of accelerating the process of replacement of temporary species by perennials, through the replacement of organic matter and possible correction of pH.

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